

# Test Results for CSTR Test 3

D. D. Lee and J. L. Collins

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Chemical Technology Division

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## CONTENTS

	<b>Page</b>
LIST OF FIGURES .....	v
LIST OF TABLES .....	vii
ACRONYMS .....	ix
ACKNOWLEDGMENTS.....	xi
ABSTRACT.....	xiii
1. INTRODUCTION .....	1
1.1 BACKGROUND .....	1
1.2 SYSTEM DESIGN .....	1
2. TEST 3 PLAN.....	3
2.1 TEST DESIGN .....	3
2.2 FEED PREPARATIONS .....	10
3. MEASUREMENTS AND SAMPLING.....	12
3.1 SAMPLING PLANS AND METHODS.....	12
3.2 MEASURING AND TEST EQUIPMENT .....	13
4. RESULTS OF TEST 3.....	13
4.1 RUN SUMMARY .....	13
4.2 DETAILS OF SYSTEM CHEMISTRY .....	16
4.2.1 Cesium Decontamination of Test 3 Salt Feed.....	16
4.2.2 Strontium Decontamination of Test 3 Salt Feed.....	17
4.2.3 Uranium Decontamination of Test 3 Salt Feed.....	19
4.2.4 HPLC Analyses for TPB Decomposition Products.....	22
4.2.5 ICP-AES Analyses for Potassium and Boron .....	22
4.2.6 Material Balances.....	25
4.3 SUBSYSTEMS OPERATION .....	25
4.4 SLURRY WASHING.....	27
5. CONCLUSIONS.....	30
6. REFERENCES.....	31



## LIST OF FIGURES

Figure		Page
1	Simplified diagram of the CSTR systems.....	2
2	Diagram of the CSTR feed systems .....	4
3	Redesigned slurry concentration filter system .....	5
4	Redesigned slurry wash filter system.....	6
5	Antifoam feed addition diagram.....	7
6	Flowsheet for the benzene monitoring system .....	8
7	Valves and piping for the benzene monitoring system.....	9
8	Foam produced from a potassium TPB slurry at an agitator speed of 1200 rpm without antifoam present.....	10
9	Test 3 <sup>137</sup> Cs DF.....	14
10	Test 3 <sup>235</sup> U DF.....	14
11	Test 3 <sup>85</sup> Sr DF.....	15
12	<sup>137</sup> Cs after CSTR 1, CSTR 2, and in the concentrate filtrate during Test 3. ....	16
13	Comparison of <sup>137</sup> Cs removal for Tests 1, 2, and 3 for CSTR 1.....	17
14	Comparison of <sup>137</sup> Cs removal for Tests 1, 2, and 3 for CSTR 2.....	18
15	Comparison of <sup>137</sup> Cs removal for Tests 1, 2, and 3 for concentrate filtrate .....	18
16	Strontium concentrations during Test 3.....	19
17	Strontium concentrations during Tests 1a and 3.....	20
18	Strontium concentrations after CSTR 1 in Tests 1a and 3.....	20
19	Strontium concentrations after CSTR 2 in Tests 1 and 3.....	21
20	Strontium concentrations in concentrate filtrate in Tests 1a and 3.....	21
21	Comparison of uranium concentrations during Test 3 .....	22
22	Comparison of uranium concentrations during Tests 1a and 3 .....	23
23	Potassium, boron, and NaTPB after CSTR 1 during Test 3.....	23

24	Potassium, boron, and NaTPB after CSTR 2 during Test 3.....	24
25	Potassium, boron, and NaTPB in concentrate filtrate during Test 3.....	24
26	Plot of the average feed rates of process fluids for Test 3.....	26
27	Plot of the cumulative feed totals for materials fed to CSTR 1 during Test 3 .....	26
28	Nitrite and TPB concentrations during Test 3 slurry washing .....	29



## LIST OF TABLES

<b>Table</b>		<b>Page</b>
1	Feed rates for CSTR feeds during Test 3.....	11
2	Average SRS salt feed preparation for Test 3.....	11
3	Simulated recycle solution preparation for Test 3.....	12
4	Process performance for Test 3.....	15
5	Results for Test 3 slurry washing at 15 mL/min.....	27
6	NaTPB recovery in the slurry washing process in CSTR tests.....	30



## ACRONYMS

CSTR	continuous-flow stirred-tank reactor
DF	decontamination factor
FTIR	Fourier transfer InfraRed analyzer
HLW	high-level waste
HPLC	high-performance liquid chromatography
ICP-AES	inductively coupled plasma–atomic emission spectroscopy
KTPB	potassium tetraphenylborate
MST	monosodium titanate
NaTPB	sodium tetraphenylborate
ORNL	Oak Ridge National Laboratory
PBs	1PB, 2PB, and 3PB decomposition products of TPB
RMAL	Radioactive Materials Analytical Laboratory
SCT	Slurry Concentrating Tank
SPP	Salt Waste Processing Program
SRS	Savannah River Site
SS	Stainless Steel
SWT	Slurry Washing Tank
TPB	tetraphenylborate
WAC	waste acceptance criteria
3PB	triphenylborane
2PB	diphenylborinic acid
1PB	phenylboronic acid



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## ABSTRACT

The goal of the Savannah River Salt Waste Processing Program (SPP) is to evaluate and select the most effective technology for the treatment of the high-level waste salt solutions currently being stored in underground storage tanks at the U.S. Department of Energy Savannah River Site (SRS) in Aiken, South Carolina. One of the three technologies currently being developed for this application is the Small-Tank Tetraphenylborate Process (STTP). This process uses sodium tetraphenylborate (NaTPB) to precipitate and remove radioactive cesium from the waste and monosodium titanate (MST) to sorb and remove radioactive strontium and actinides. Oak Ridge National Laboratory is demonstrating this process at the 1:4000 scale using a 20-L capacity continuous-flow stirred-tank reactor (CSTR) system. Since March 1999, three operating campaigns of the 20-L CSTR have been conducted. The ultimate goal is to verify that this process, under certain extremes of operating conditions, can meet the minimum treatment criteria necessary for processing and disposal at the Savannah River Saltstone Facility. The waste acceptance criteria (WAC) for  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$ , and total actinides are  $<40$  nCi/g,  $<40$  nCi/g, and  $<18$  nCi/g, respectively. However, to allow for changes in process conditions, SPP is seeking a level of treatment that is about 50% of the WAC. The bounding separation goals for  $^{137}\text{Cs}$  and  $^{90}\text{Sr}$  are to obtain decontamination factors (DFs) of 40,000 (99.998% removal) and 26 (96.15% removal), respectively. The DF is defined as the concentration of contaminant in the waste divided by the concentration of contaminant in the effluent stream.

Tests 1 and 2 were conducted in June and July 1999, and the results are documented in ORNL/TM-1999/234. These initial tests, using simulants traced with radioactive materials, verified that the STTP process could achieve the necessary cesium, strontium, and actinide decontamination under standard operating conditions, with and without the recovery and recycle of the excess sodium tetraphenylborate (NaTPB) that was not precipitated with potassium and cesium. TPB decomposition, the major side reaction of the process, did not occur in Test 2 despite the addition of materials (synthetic sludge and modified Enhanced Comprehensive Catalyst) that were known to cause TPB to degrade.

Test 3 was a 72-h demonstration that was conducted August 2000. The objectives were to examine the effectiveness of an improved antifoam (IITB52) in minimizing foam formation and to determine if its presence had a damaging effect on the removal of  $^{137}\text{Cs}$ ,  $^{85}\text{Sr}$ , and  $^{235}\text{U}$  from simulated waste. To more aggressively test the antifoam, the slurry in each vessel was mixed at 1200 to 1250 rpm while maintaining the temperature at  $25^{\circ}\text{C}$ . This rate of mixing was found to cause heavy foam formation in a control test that was conducted prior to Test 3. No evidence of excessive foam formation was noted in Test 3. Excessive foam formation would have been indicated by the presence of foam in the analytical samples of slurries taken, by the presence of foam in the clear plastic slurry transfer lines between the vessels, by significant changes in the level probe readings in process vessels, and by large changes in the slurry density measurements in the slurry concentration system.

The decontamination goal of obtaining a decontamination factor (DF) of 40,000 or 99.998% removal was exceeded in all process vessels. A  $\text{DF}_{\text{Cs}}$  of 40,000 was obtained in 14 h in CSTR 1 and afterward ranged between 41,000 and 270,000. It took about 32 h for the DF to reach 40,000 in CSTR 2, and for the remainder of the test, the DF ranged from 49,000 to 77,000. A  $\text{DF}_{\text{Cs}}$  of 37,000 was obtained in the slurry concentrating tank (SCT) in about 32 h. It subsequently fluctuated between 21,000 and 59,000 and was the highest at the end of the test.

No impact of the IITB52 on the sorption of strontium and uranium by the MST was noted. A  $\text{DF}_{\text{Sr}}$  of 26 was needed to reach the WAC standard for saltstone. A DF of 30 was obtained for strontium in the SCT in about 24 h, and near the end of the test, it was 100. A DF of 5 for uranium in the SCT was obtained at the end of the test. This is indicative that the other actinides neptunium and plutonium would have DFs that would have met the WAC for saltstone.

At the end of Test 3, the concentrated slurry in the concentrating system was transferred to the Slurry Washing Tank and washed with deionized water to recover unutilized NaTPB. IITB52 was added to the tank at a rate of 0.00312 mL/min. Because of the method of starting Test 3 in which the process

vessels were first filled with the 4.7 M Na<sup>+</sup> salt feed, a large fraction of the TPB added was precipitated as potassium tetraphenylborate (KTPB), leaving only 0.33 mol of excess TPB in the slurry concentrating system available for recovery during water washing. The collected washwater had a concentration of 0.0033 mol/L TPB (0.095 mol TPB total), which was 29% of the available TPB. This amount was greater than the amount recovered in Test 1a, in which only 12% of the available TPB was recovered and no IITB52 was used. This indicates that the antifoam did not interfere with the recovery of the precipitated NaTPB in this case. The higher mixing rate in Test 3 (1200 vs 600 rpm) may have aided in the larger fraction of TPB being recovered.



## TEST RESULTS FOR CSTR TEST 3

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### 1. INTRODUCTION

#### 1.1 BACKGROUND

The Small-Tank Tetraphenylborate Process (STTP) is one of three separations processes currently being examined by the Savannah River High-Level Waste (HLW) Salt Waste Processing Program (SPP) as an alternative to the In-Tank Precipitation Process for treatment of highly radioactive Savannah River Site (SRS) tank waste. The ultimate goal of the process is to decontaminate the SRS salt waste so that it meets the waste acceptance criteria (WAC) for processing and disposal at the Savannah River Saltstone Facility. The WAC for  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$ , and alpha are  $<40$  nCi/g,  $<40$  nCi/g and  $<18$  nCi/g, respectively. However, to allow for changes in process conditions, the SPP is seeking a level of treatment that is about 50% of the WAC. Supernatants in high-OH SRS tanks that are 6.4 M in  $\text{Na}^+$  have cesium concentrations in the range of 0.000415 M. Cesium concentrations for average SRS waste with the same  $\text{Na}^+$  concentration are in the range of 0.000157 M. The  $^{137}\text{Cs}$  radioactivity level for these wastes are 1.1 and 0.41 Ci/L, respectively.

A 1:4000-scale, 20-L continuous-flow stirred-tank reactor (CSTR) system was designed and fabricated at Oak Ridge National Laboratory (ORNL) for development of this process. The goal of the 20-L CSTR test program is to evaluate the performance of the 20-L CSTR system under conditions that simulate the anticipated full-scale operation of the system. The basic chemistry of the process makes use of tetraphenylborate (TPB) to remove cesium by precipitation and monosodium titanate (MST) to remove strontium and actinides (plutonium, neptunium, and uranium) by sorption. The CSTR system was to be operated in a contained and shielded hot cell environment to allow the use of radioactive materials. Testing of the CSTR system would determine if this approach could (1) provide the cesium, strontium, and actinide separations needed for treatment of the SRS salt waste; (2) perform satisfactorily with an improved antifoam agent, and (3) perform satisfactorily while active decomposition of TPB was taking place with and without the recovery and recycle of unreacted TPB.

Test 1a (Tests 1b and 1c were not needed) and Test 2 were conducted in FY 1999 and met performance objectives in verifying that the required separations of cesium, strontium, and actinides could be provided by this process.<sup>1</sup>

The primary goal of Test 3 was to examine the effectiveness of an improved antifoam (IITB52) in minimizing foam formation and to determine if its presence in the TPB slurry had significant impact on cesium, strontium, and actinide removal performance. Test 3 was planned to be a single-pass, 80-h test that provided about ten volume turnovers in the CSTRs. No sludge or catalyst was added to the salt feed, and the slurry solutions were maintained at  $\sim 25^\circ\text{C}$  during the test.<sup>2</sup>

#### 1.2 SYSTEM DESIGN

The apparatus used in the test system was sized and designed to provide reliable data for scaleup to larger systems. Figure 1 shows a simplified process flowsheet for the system. All feeds are introduced into CSTR 1, and the degree of decontamination of the process stream in any vessel is always defined based on the feed to this reactor. Most of the precipitation of cesium takes place in CSTR 1. CSTR 2 provides additional residence time for a small amount of additional cesium precipitation to occur. However, the primary role of CSTR 2 in the process is to provide the required additional residence time needed for the MST to sorb the actinides present in the real waste. The test system, which is located in Hot Cell B, Building 4501, includes two CSTRs in series, each with a 15-L working volume; a slurry

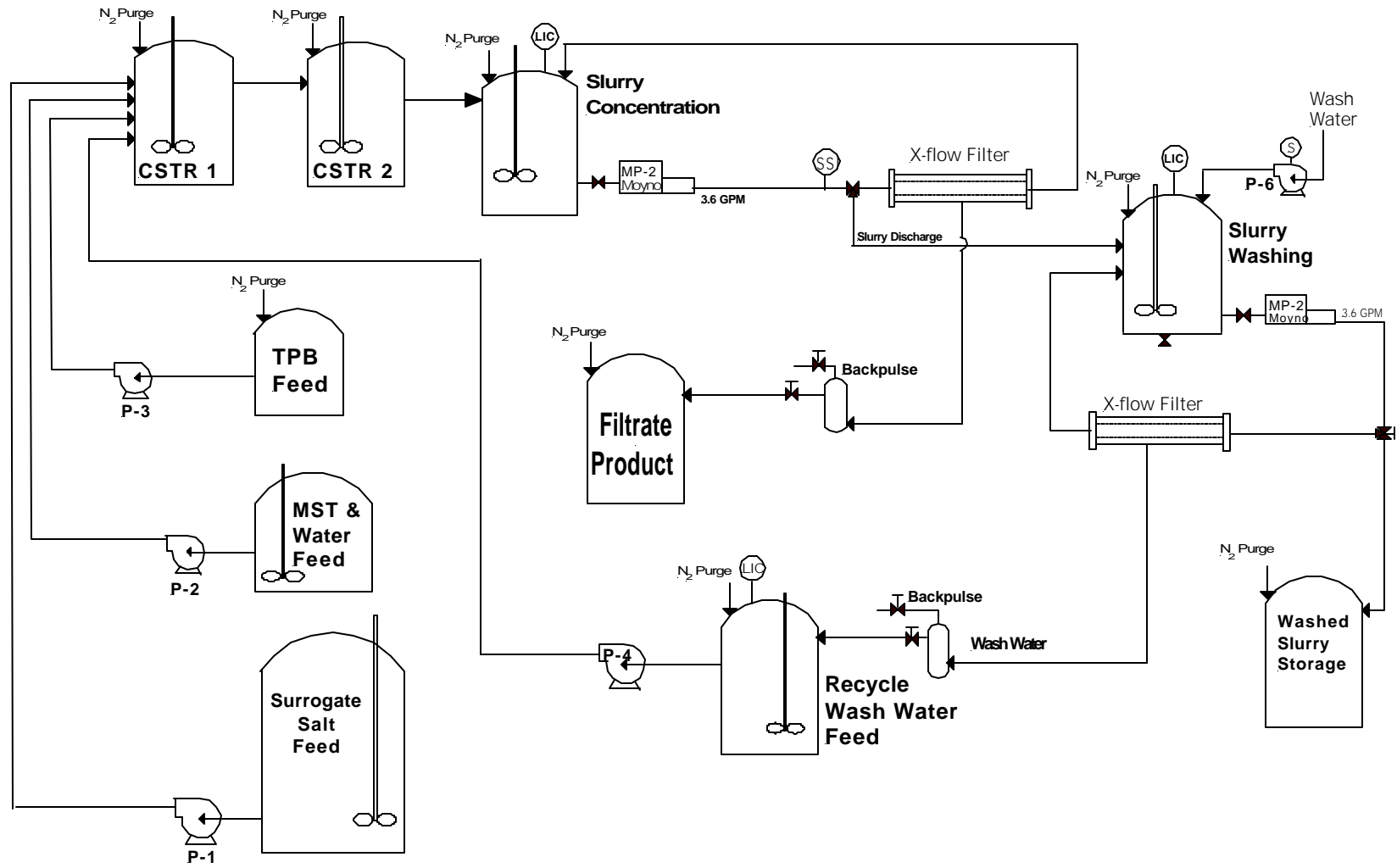


Fig. 1. Simplified diagram of the CSTR systems.

concentration tank with a cross-flow filter to concentrate the slurry; a concentrated-slurry washing tank with a cross-flow filter; and various tanks and pumps to integrate the operation. Two 55-gal feed tanks with mixing systems were located in Hot Cell A.

After Test 2 was concluded, plans were developed for making needed improvements to the CSTR system. Modifications were needed to provide improved control systems and data collection, vessel off-gas monitoring for benzene, improved level monitoring, and improved antifoam feed capability. In more detail, the CSTR modification in Cell B involved the following:

1. Rerouting some of the stainless steel piping to change the flow configuration.
2. Installing eight electric valves for improved cross-flow filter backpulse systems. (Six were new, and two replaced manual valves.)
3. Replacing two three-way manual valves with five-way electric valves to improve the ability to direct slurry flow.
4. Installing four pressure transducers and two flow transmitters (part of the improved backpulse system).
5. Installing/retrofitting electrical/instrument pigtail connections on cables for valves and instruments (for ease of connection and removal during cell entries) and installing signal cables for new valves and instruments (the four pressure transducers and two flow transmitters).
6. Installing new flex tubing and replacing some flex tubing with hard-piped stainless steel (SS) tubing on the CSTRs and filter systems.
7. Installing a benzene monitoring system to monitor the off-gas from CSTR 1, CSTR 2, the Slurry Concentrating Tank (SCT), and the Slurry Wash Tank (SWT). This included 4 GASMET Fourier transform infrared (FTIR) analyzers; 4 pumps; 8 mass-flow meters; 12 solenoid valves with appropriate plumbing; 4 liquid traps; and necessary tubing, heat tracing, thermocouples, and electrical wiring.
8. Installing new antifoam delivery systems to the two CSTRs and to the SCT and SWT.
9. Installing a new level probe in the concentrate tank identical to the one in the SWT and a new level probe in CSTR 1 (which can be used as a backup for the either of the other two level probes).
10. Transferring the simulated salt feed and washed slurry concentrate equipment from Hot Cell A to Hot Cell C.

Figures 2–7 provide the updated flowsheets for the TPB process test system in Hot Cell B. Figure 2 shows the CSTR feed systems; Fig. 3, the slurry concentrate system; and Fig. 4, the slurry wash system. Figure 5 shows the antifoam feed system, and Figs. 6 and 7 show the CSTR benzene monitoring flowsheets.

## 2. TEST 3 PLAN

### 2.1 TEST DESIGN

The primary goal of Test 3 was to examine the effectiveness of an improved antifoam (IITB52) in minimizing foam formation and to determine if its presence in the TPB slurry had a deleterious effect on the decontamination factors (DFs) of cesium, strontium and uranium. Test 3 was planned to be a single-pass (no recycle of recovered TPB), 80-h test that provided about ten volume turnovers in the CSTRs. No sludge or catalyst was added to the salt feed, and the slurry solutions were maintained at ~25°C during the test. IITB52 is a water-soluble liquid with a density of 1.01 g/mL. It was pumped continuously with syringe pumps to the CSTRs and SCT to limit foam formation. The concentration of IITB52 in the salt feed slurry was maintained at 50 ppmv (parts per million by volume) in the CSTRs and 100 ppmv in the SCT. The CSTRs were filled with diluted salt feed (4.7 M Na<sup>+</sup>) at the start of the test, and about 3 L was added to the SCT.

An agitator speed of 1200 rpm was chosen for the test based on the results of a preliminary test performed with a slurry of KTPB. A photo of the foam generated during this test is shown in Fig. 8.

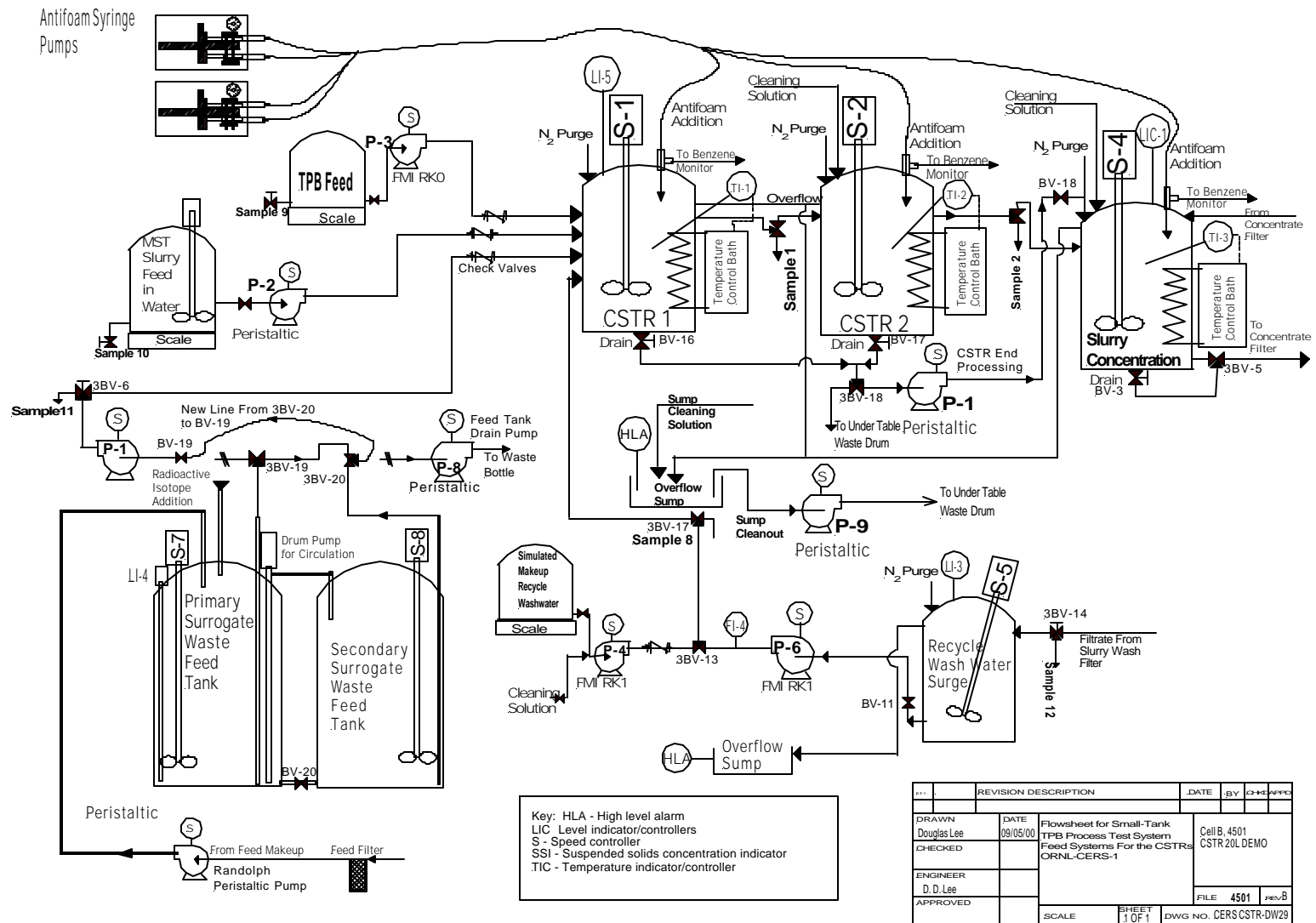


Fig. 2. Diagram of the CSTR feed systems.

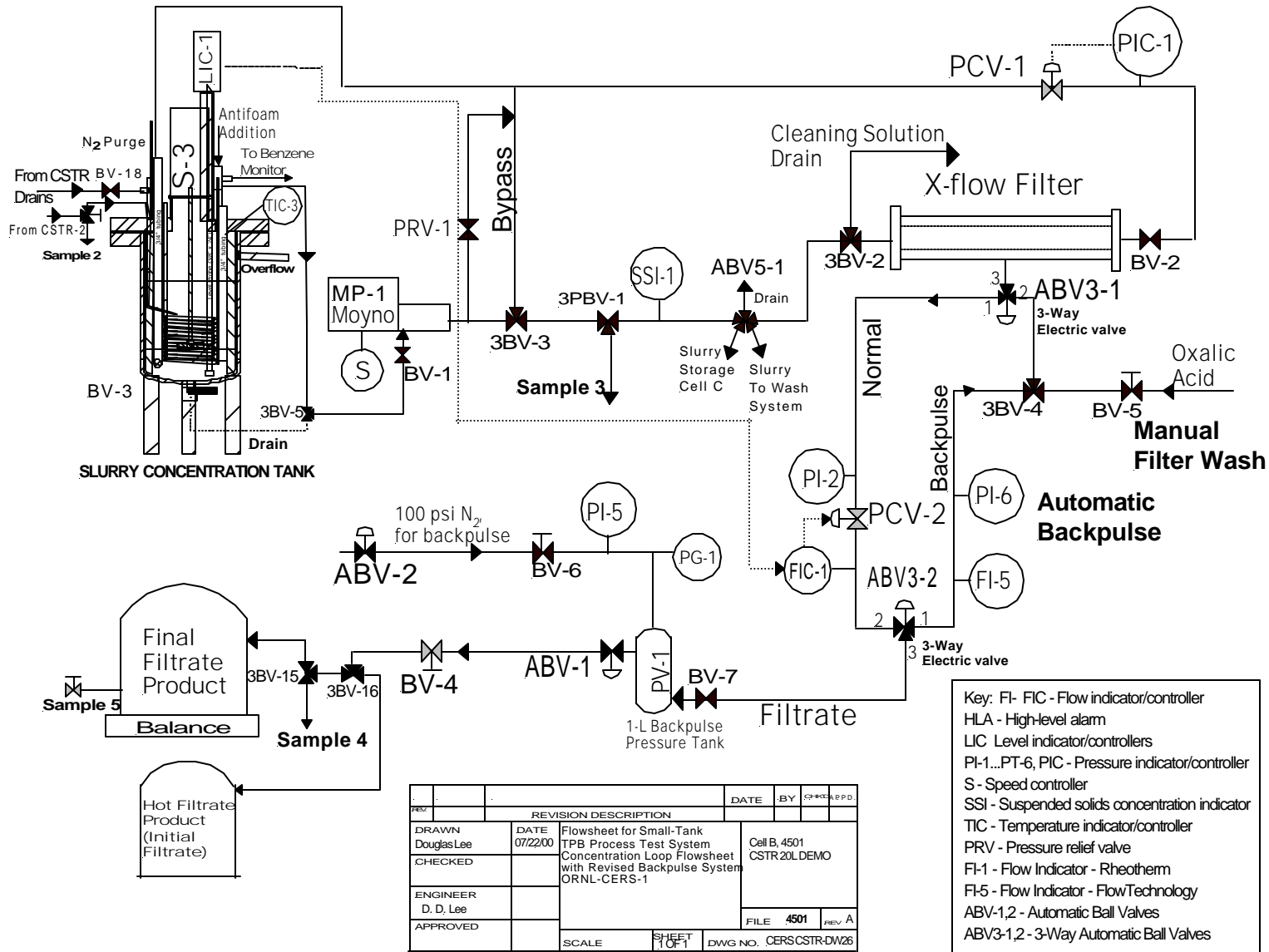


Fig. 3. Redesigned slurry concentration filter system.

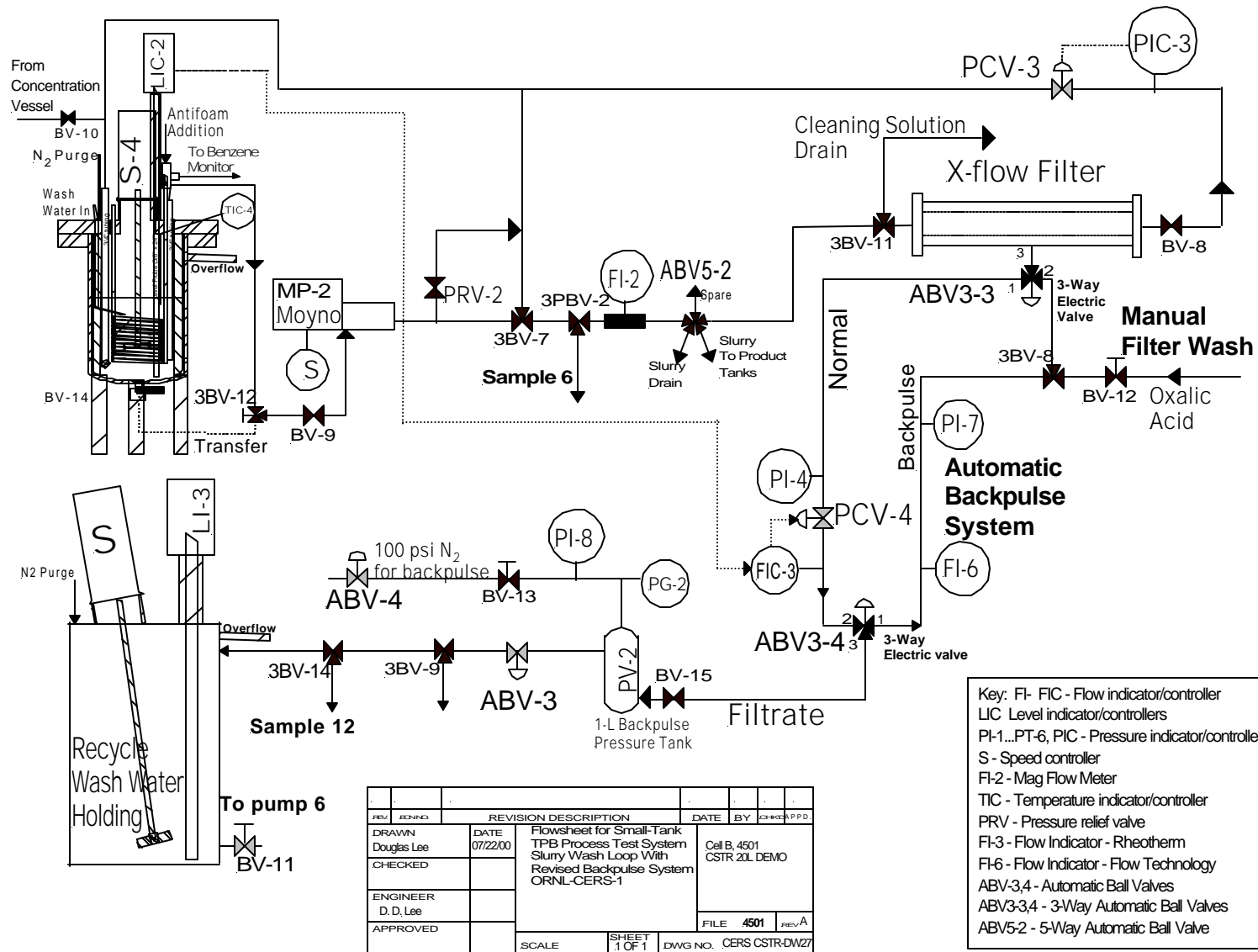
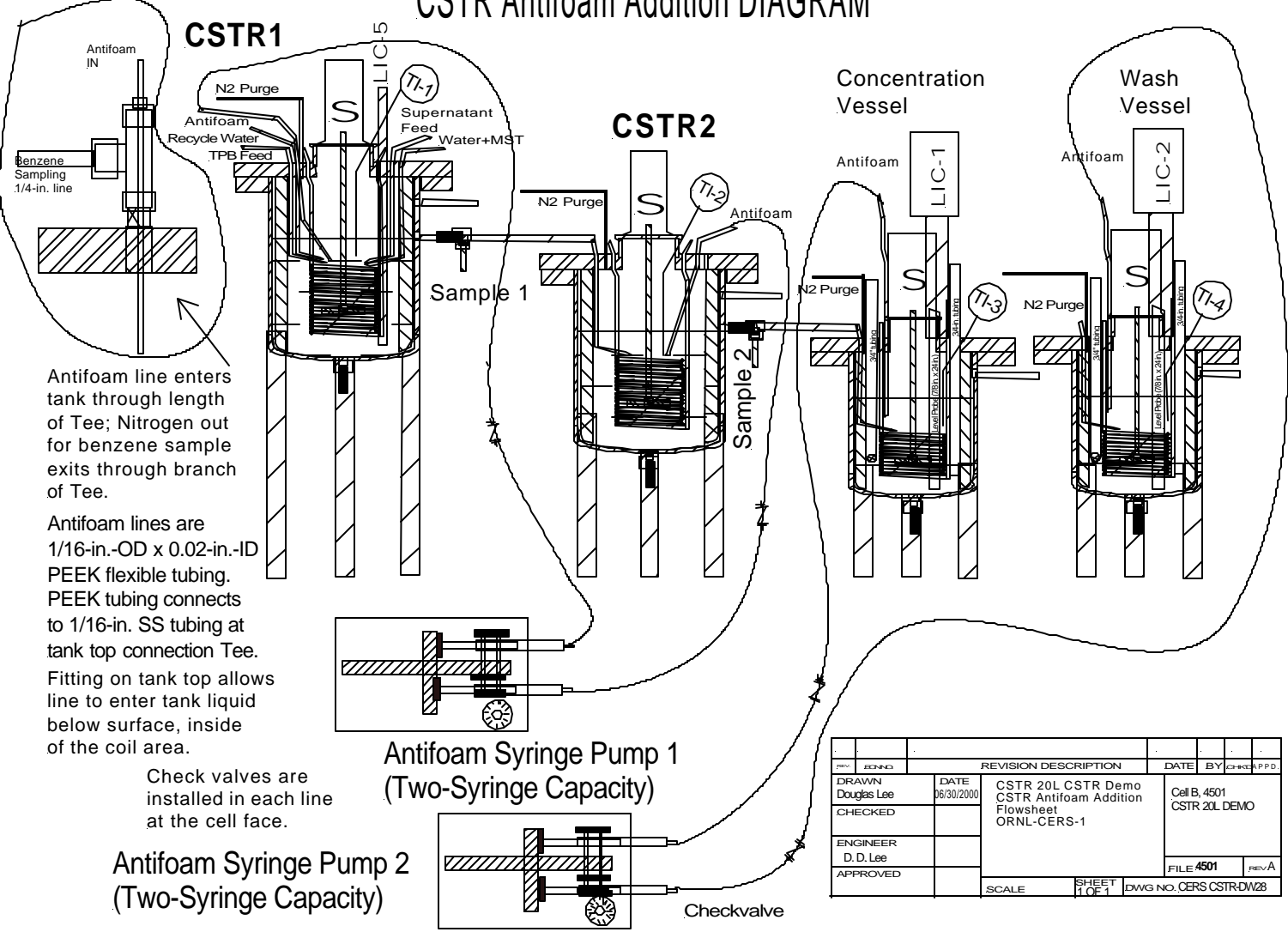


Fig. 4. Redesigned slurry wash filter system.

# CSTR Antifoam Addition DIAGRAM



Antifoam line enters tank through length of Tee; Nitrogen out for benzene sample exits through branch of Tee.

Antifoam lines are 1/16-in.-OD x 0.02-in.-ID PEEK flexible tubing. PEEK tubing connects to 1/16-in. SS tubing at tank top connection Tee. Fitting on tank top allows line to enter tank liquid below surface, inside of the coil area.

Check valves are installed in each line at the cell face.

Antifoam Syringe Pump 2 (Two-Syringe Capacity)

Antifoam Syringe Pump 1 (Two-Syringe Capacity)

Checkvalve

REV	BY	DATE	REVISION DESCRIPTION	DATE	BY	CHK	APP	P.P.D.
DRAWN	Douglas Lee	06/30/2000	CSTR 20L CSTR Demo CSTR Antifoam Addition Flowsheet ORNL-CERS-1					
CHECKED								
ENGINEER	D. D. Lee							
APPROVED								
			SCALE	SHEET	FILE			
			1 OF 1	4501				
			DWG NO. CERS CSTRD/W28					

Fig. 5. Antifoam feed addition diagram.

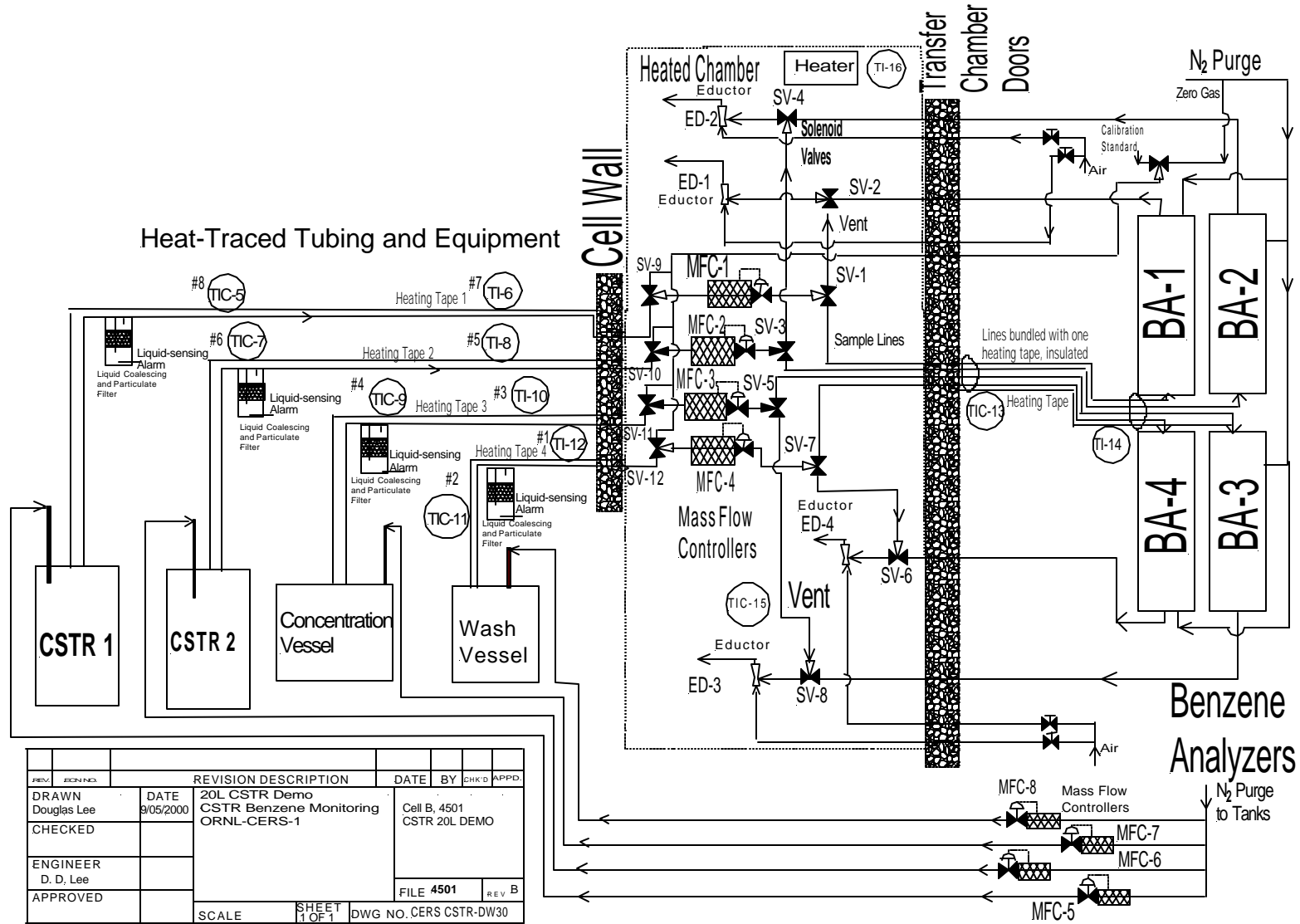


Fig. 6. Flowsheet for the benzene monitoring system.



6

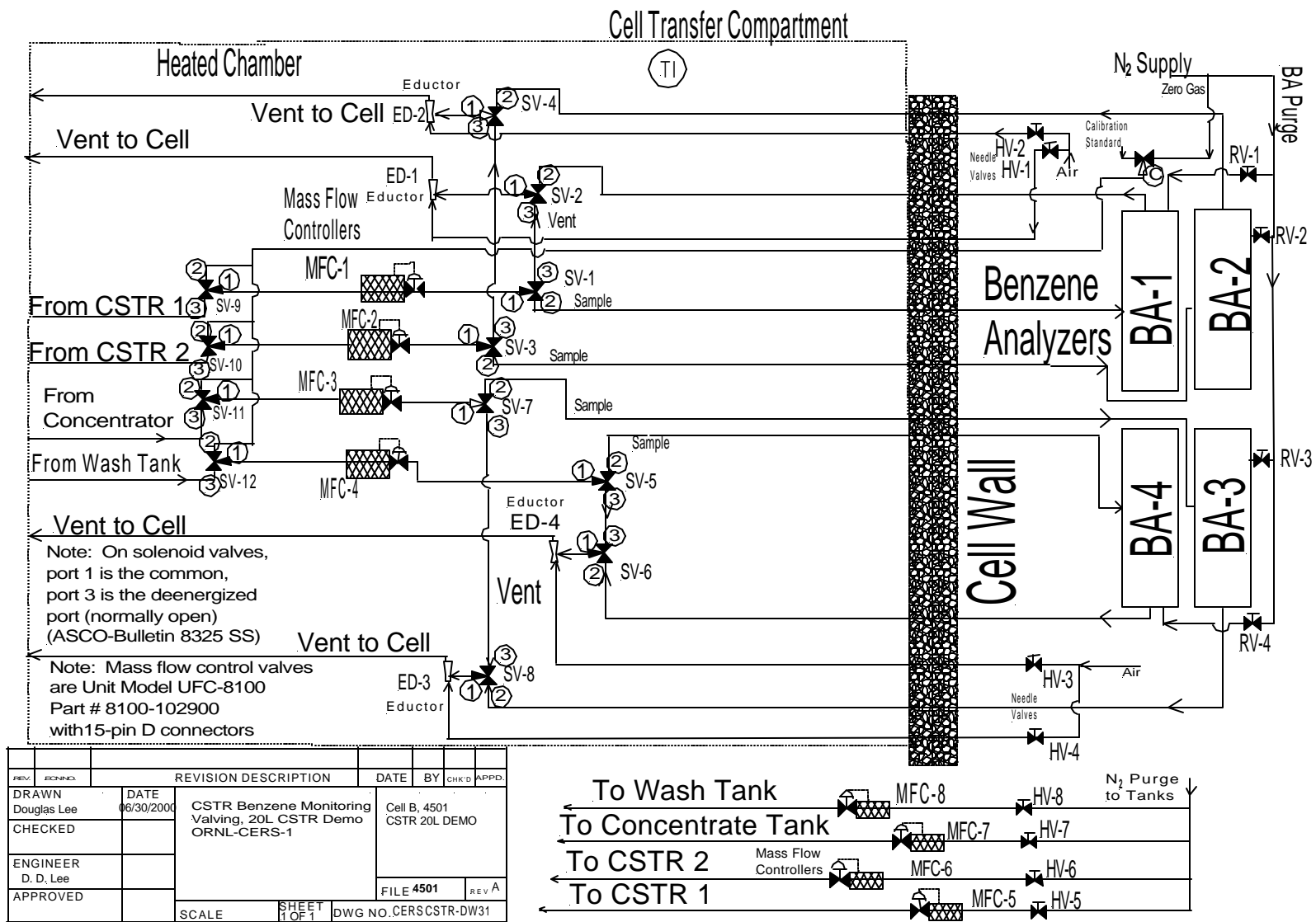


Fig. 7. Valves and piping for the benzene monitoring system.



**Fig. 8. Foam produced from a potassium TPB slurry at an agitator speed of 1200 rpm without antifoam present. (Normal operations at 600 rpm produced very little foam.)**

## 2.2 FEED PREPARATIONS

The procedures and materials for feed preparation are given in CERS/SR/TPB/005.<sup>3</sup> Table 1 shows the feed rates for the test, and Table 2 gives the masses of SRS average salt feed compounds needed to prepare the 140 L of salt feed for the test. Table 3 gives the masses of the chemical compounds needed to prepare 37 L of synthetic recycle wash water. This was added to CSTR 1 along with the other feeds to simulate the recycle of wash water from the TPB recovery and recycle process. The quantity of synthetic wash water prepared assumed a 90-h run time at 6.86 mL/min. The synthetic recycle water contained enough sodium tetraphenylborate (NaTPB), 0.03303 M, to provide 60% excess of the stoichiometric NaTPB requirement to CSTR 1. To provide a stoichiometric amount of NaTPB to CSTR 1 to precipitate the potassium, 6.52 L of the NaTPB feed was needed (0.68 mL/min feed rate). The concentrations of NaTPB and NaOH in this feed solution were 0.5545 and 0.96 M, respectively. A total of 1237.2 g NaTPB [(0.5545 mol/L)(6.52 L)(342.23 g/mol)] and 25.05 g NaOH [(0.096 mol/L)(6.52 L)(40 g/mol)] were needed for the preparation. The concentration of MST (preparation is given in reference 3) in the slurry solution in CSTR 1 was set at about 0.5 g/L. The concentration of <sup>137</sup>Cs in the salt feed (average SRS recipe) was ~7.2 mCi/L, which was diluted by other feeds to ~5.5 mCi/L of slurry in the CSTRs. The concentration of <sup>85</sup>Sr and uranium (93 wt % <sup>235</sup>U) in the salt in the salt feed was 0.026 mCi/L and 0.8 ppm, respectively.

The CSTRs were filled prior to testing with salt feed and deionized water at a rate of 21.5 mL/min salt feed and 7.78 mL/min deionized water, providing a Na<sup>+</sup> concentration of 4.7 M. About 3 L of the diluted salt feed was added to the SCT after the CSTRs were filled. IITB52 antifoam was also added at a rate to provide concentrations of 50 ppm in the CSTRs and 100 ppm in the SCT. During the start-up phase, the CSTRs each received 0.77 mL of antifoam, and the SCT received 0.15 mL of antifoam. In addition, each 30-ft antifoam line has a 2.7 mL capacity, so 2.7 mL was pre-fed through each line before starting to add the required volumes to the tanks.

**Table 1. Feed rates for CSTR feeds during Test 3**

Feed materials	Feed rates		
		Planned	Actual
	(mL/min)	(mL/h)	(mL/min)
Salt feed	21.5	1209	21.50
Recycle wash water with excess NaTPB	6.86	411.6	7.35
MST + water	2.14	128.4	2.08
NaTPB	0.68	40.8	0.71
IITB52 feed to CSTR 1 and CSTR 2	0.00156	0.0935	0.00156
IITB52 feed to SCT	0.00312	0.1871	0.00312
Total feed rate	31.19		31.64

**Table 2. Average SRS salt feed preparation for Test 3**

Chemical Specie	Molarity (mol/L)	Compound	Molecular weight (g/mol)	Mass (g/L)	Test 3 mass <sup>a,b</sup> (g/151.4 L)
Cs <sup>-c</sup>	0.00016	CsCl	168.36	0.027	4.1
K <sup>-</sup>	0.0171	KNO <sub>3</sub>	101.1	1.729	261.8
OH <sup>-</sup>	3.611	NaOH	40.00	144.44	21,868.2
NO <sub>3</sub> <sup>-</sup>	1.365 <sup>d</sup>	NaNO <sub>3</sub>	84.99	116.01	17,563.9
NO <sub>2</sub> <sup>-</sup>	0.594	NaNO <sub>2</sub>	69.00	40.99	6,205.9
AlO <sub>2</sub> <sup>-</sup>	0.354	Al(NO <sub>3</sub> ) <sub>3</sub> 9H <sub>2</sub> O	375.14	132.80	20,105.9
CO <sub>3</sub> <sup>2-</sup>	0.183	Na <sub>2</sub> CO <sub>3</sub> H <sub>2</sub> O	124.01	22.69	3,435.3
SO <sub>4</sub> <sup>2-</sup>	0.171	Na <sub>2</sub> SO <sub>4</sub>	142.04	24.29	3,677.5
Cl <sup>-</sup>	0.0286	NaCl	58.4	1.670	252.9
F <sup>-</sup>	0.0366	NaF	41.99	1.537	232.7
HPO <sub>4</sub> <sup>2-</sup>	0.0114	Na <sub>2</sub> HPO <sub>4</sub> 7H <sub>2</sub> O	268.09	3.056	462.7
C <sub>2</sub> O <sub>4</sub> <sup>2-</sup>	0.0091	Na <sub>2</sub> C <sub>2</sub> O <sub>4</sub>	134.00	1.219	184.6
SiO <sub>3</sub> <sup>2-</sup>	0.0046	Na <sub>2</sub> SiO <sub>3</sub> 9H <sub>2</sub> O	284.2	1.307	197.9
MoO <sub>4</sub> <sup>2-</sup>	0.00023	Na <sub>2</sub> MoO <sub>4</sub> 2H <sub>2</sub> O	241.95	0.056	8.5
Sr <sup>2+e</sup>		Sr(NO <sub>3</sub> ) <sub>2</sub>	211.75	0.000011	0.0167
Uranium <sup>e</sup>		UO <sub>2</sub> (NO <sub>3</sub> ) <sub>2</sub>	391.13	0.00154	0.2332
Na <sup>+</sup>	6.394				
H <sub>2</sub> O				796.	120,514
<b>Totals</b>				<b>1,289.8</b>	<b>195,276</b>

<sup>a</sup>Assumes 90-h run time at 21.5 mL (salt feed)/min, or 0.412 L (salt feed) /h; also accounts for salt feed added to CSTRs and SCT at beginning of test.

<sup>b</sup>Calculated density = (195,276 g)/(151,400 mL) = 1.29 g/mL (140 L of salt feed was actually added to feed tanks in Hot Cell C and the tracers were added with good mixing; measured density was 1.281 g/mL).

<sup>c</sup>4.21 g of Cs Cl (3.32 g Cs) in 70 mL water was mixed well with 1.35 Ci<sup>137</sup> tracer (mass of Ci<sup>137</sup> was 0.016; but total cesium mass in tracer was 0.073 g). The total mass of <sup>133</sup>Cs and <sup>137</sup>Cs used in Test 3 was 3.393 g, which provided the needed 0.000167 M cesium in the salt feed. The <sup>137</sup>Cs concentration in the salt feed was 9.1 mCi/L and 6.3 mCi/L in the 4.7 M Na<sup>+</sup> slurry.

<sup>d</sup>The target molarity for NO<sub>3</sub><sup>-</sup> is 2.45 mol/L. The addition of Al(NO<sub>3</sub>)<sub>3</sub> 9H<sub>2</sub>O and KNO<sub>3</sub> provides 1.062 and 0.017 mol NO<sub>3</sub><sup>-</sup> /L, respectively. [2.45 - (1.068 + 0.017)] = 1.365 mol (NaNO<sub>3</sub>)/L extra needed].

<sup>e</sup>Strontium-85 was added to obtain a feed concentration of (0.068 mCi/L) and cold strontium was added to obtain a concentration of 0.047 mg/L, and uranium (93% <sup>235</sup>U) was added to obtain a concentration of 0.93 mg/L. The calculated strontium and uranium concentration for the 4.7 M Na<sup>+</sup> slurry were 0.032 mg/L (0.046 mCi/L) and 0.64 mg/L, respectively. Details about the preparations of the strontium and uranium stock solutions used can be found in ref. 3.

**Table 3. Simulated recycle solution preparation for Test 3**

Chemical specie	Molarity (mol/L)	Compound	Molecular weight (g/mol)	Mass Fraction (g/L)	Total mass <sup>a,b</sup> (g/37 L)
TPB <sup>-</sup>	0.03303	NaTPB	342.23	11.30	418.10
OH <sup>-</sup>	0.6584	NaOH	40.00	26.346	974.80
NO <sub>3</sub> <sup>-</sup>	0.2531	NaNO <sub>3</sub>	84.99	21.248	786.18
NO <sub>2</sub> <sup>-</sup>	0.1083	NaNO <sub>2</sub>	69.00	7.473	276.50
AlO <sub>2</sub> <sup>-</sup>	0.0645	Al(NO <sub>3</sub> ) <sub>3</sub> 9H <sub>2</sub> O	375.14	24.196	895.25
CO <sub>3</sub> <sup>2-</sup>	0.0334	Na <sub>2</sub> CO <sub>3</sub> H <sub>2</sub> O	124.01	4.142	153.25
SO <sub>4</sub> <sup>2-</sup>	0.0312	Na <sub>2</sub> SO <sub>4</sub>	142.04	4.432	163.98
Cl <sup>-</sup>	0.0052	NaCl	58.4	0.304	11.25
F <sup>-</sup>	0.0067	NaF	41.99	0.281	10.40
HPO <sub>4</sub> <sup>2-</sup>	0.0015	Na <sub>2</sub> HPO <sub>4</sub> 7H <sub>2</sub> O	268.09	0.563	25.33
C <sub>2</sub> O <sub>4</sub> <sup>2-</sup>	0.0017	Na <sub>2</sub> C <sub>2</sub> O <sub>4</sub>	134.00	0.228	8.44
SiO <sub>3</sub> <sup>2-</sup>	0.0008	Na <sub>2</sub> SiO <sub>3</sub> 9H <sub>2</sub> O	284.2	0.227	8.40
MoO <sub>4</sub> <sup>2-</sup>	0.00004	Na <sub>2</sub> MoO <sub>4</sub> 2H <sub>2</sub> O	241.95	0.0097	0.36
Na <sup>+</sup>	1.20				
H <sub>2</sub> O				950	35,150
<b>Totals</b>				<b>1,050.75</b>	<b>38,882.2</b>

<sup>a</sup> Assumes 90-h run time at 6.86 mL/min (or 0.412 L/h).

<sup>b</sup> Calculated density for preparation = (38,881.5 g)/(37,000 mL) = 1.057 g/mL.

### 3. MEASUREMENTS AND SAMPLING

#### 3.1 SAMPLING PLANS AND METHODS

During the tests, samples were taken for analysis according to the sampling plan given in CERS/SR/TPB/006, Rev. 1, *Sampling Plan and Procedures*.<sup>4</sup> After preparation, the salt feed solution was analyzed for K, <sup>137</sup>Cs, and <sup>85</sup>Sr. Effluent samples were taken for each CSTR (sample ports 1 and 2), and a filtrate sample was taken from the SCT filtration system (sample port 4) every 4 h (samples of 55–65 mL each, collected over a 2min duration). Each sample was analyzed using an on-site gamma counter for <sup>137</sup>Cs and <sup>85</sup>Sr. The counter results indicated an error margin of ±10% and minimum detectable levels of 3E-5 µCi/g for both <sup>85</sup>Sr and <sup>137</sup>Cs. At 8-h time intervals, duplicate samples were analyzed by the ORNL Radioactive Materials Analysis Laboratory (RMAL) for <sup>137</sup>Cs (±10% error margin), <sup>85</sup>Sr (±10%), <sup>235</sup>U (±10%), K (±10%), B (±10%), and NaTPB (±20%). Every fifth sample (20-h interval) was analyzed by Reverse-Phase High-Performance Liquid Chromatography (HPLC) for TPB decomposition products. Portions of some of the samples were analyzed for uranium by delayed neutron counting (±10% error margin). The slurry samples from the CSTRs were also examined for the presence of foam. Other samples were archived in case additional data were needed.

Samples from sample ports 1 (SP-1) and 2 (SP-2) were prepared for analysis by filtering the sample through either a 0.45-µm filter or a 0.80/0.22-µm double filter and placing the filtrate in new sample bottles for transfer to RMAL. Some of the samples were very difficult to filter with the 0.45-µm filter, which necessitated use of the double filter. The double-filter design facilitated easier and faster filtration of the sample, which reduced personnel radiation dose and the potential for contamination issues. Analytical results for the same sample, filtered using both methods, were comparable. HPLC was used for NaTPB and TPB decomposition products, including triphenylborane (3PB), diphenylborinic

acid (2PB), phenylboronic acid, and phenol. Official gamma counting by RMAL was performed using either Canberra or ORTEC GMX closed-end coaxial detectors for  $^{137}\text{Cs}$  and strontium ( $^{85}\text{Sr}$ ) with an error margin of  $\pm 10\%$ . Inductively coupled plasma Atomic Emission Spectroscopy (ICP-AES) was used for potassium and boron analysis. Prior to analysis, potassium and boron samples were digested in nitric acid in a microwave oven. The filtrate samples from SCT (sample port 4) were sent, as collected, to RMAL for analysis after a small sample (1–3 mL) was placed in a counting tube for on-site  $^{137}\text{Cs}$  and  $^{85}\text{Sr}$  counting. All samples destined for RMAL were placed in a refrigerator (2–4°C) after preparation, until they could be transported. A chain-of-custody procedure in accordance with the Sample Management Office was followed for all samples sent to RMAL.

### 3.2 MEASURING AND TEST EQUIPMENT

Electronic balances, which are calibrated annually by ORNL Plant and Equipment (P&E) personnel, were used for solids and liquid mass measurements in preparing the simulants. Class A volumetric flasks and graduated cylinders were used for simulant preparation. For on-site monitoring of  $^{137}\text{Cs}$  and  $^{85}\text{Sr}$  during the CSTR tests, a gamma-counting system consisting of an ORTEC model GMX-45220-P-S intrinsic germanium detector, an ORTEC model 672 counting system amplifier, a Canberra Accuspec-A MCA card, and Canberra Genie-2000 spectroscopy software was used.

Instrumentation used in the CSTR system was calibrated before startup. Calibration records are maintained in registered logbooks and in a controlled project file. A revised list of instruments and corresponding calibration schedules was generated after the modifications to the CSTR system had been reviewed and approved by the SPP and Oak Ridge investigators.<sup>5</sup>

## 4. RESULTS OF TEST 3

### 4.1 RUN SUMMARY

Test 3 was successfully conducted August 14–18, 2000, with no unscheduled downtime. During the operation, a total of about 93 L (24.5 gal) of radioactive simulated SRS salt feed was treated with ~135 L (~36 gal) of decontaminated filtrate and ~6.5 L of concentrated TPB/MST slurry (>9 wt % insoluble solids) collected. Overall, the DF values for cesium, strontium, and uranium exceeded the WAC standards needed for filtrate disposal in saltstone. The concentration of the NaTPB in the recycle water provided 60% excess NaTPB to ensure the needed cesium DF of 40,000. Figure 9 shows the  $^{137}\text{Cs}$ - DFs versus time profiles for CSTR 1, CSTR 2, and the Slurry Concentration Vessel. A DF of 40,000 was obtained in CSTR 1 about 12 h after the initiation of NaTPB feed. This DF was also achieved in CSTR 2 after about 26 h. In the Slurry Concentration Tank, the DF for cesium reached 10,000 in about 36 h and 40,000 in about 70 h. Periods of 12, 18, and 22 h, respectively, were required to obtain a DF of 2 for uranium in CSTR 1, CSTR 2, and the SCT. A DF of about 5 for uranium was obtained at the end of the test in the concentrate filtrate. As shown in Fig. 10, the loss of DF for uranium was very noticeable for CSTR 1 in the 64- to 72-h samples and for CSTR 2 in the 68- and 72-h samples as the concentration increased. During this period, the MST feed line to CSTR 1 became disconnected at a location not visible to operating personnel. After about 6 h, the disconnected line was discovered and repaired. Figure 11 shows the DF versus time profile for  $^{85}\text{Sr}$ . After 36 h, a strontium DF of 30 was obtained in the SCT, and the DF slowly increased to about 50 at the end of the test.

The overall system operation was performed in accordance with the test plans, and the behavior of the system components and overall system performance satisfied the test objectives. No foam was detected in any slurry samples from CSTR 1 or CSTR 2. No foam was visible in the slurry transfer lines between CSTR 1 and CSTR 2, between CSTR 2 and the SCT, and no other problems with slurry transfer occurred. The most difficult operational problems in Test 3 occurred with the MST/dilution water

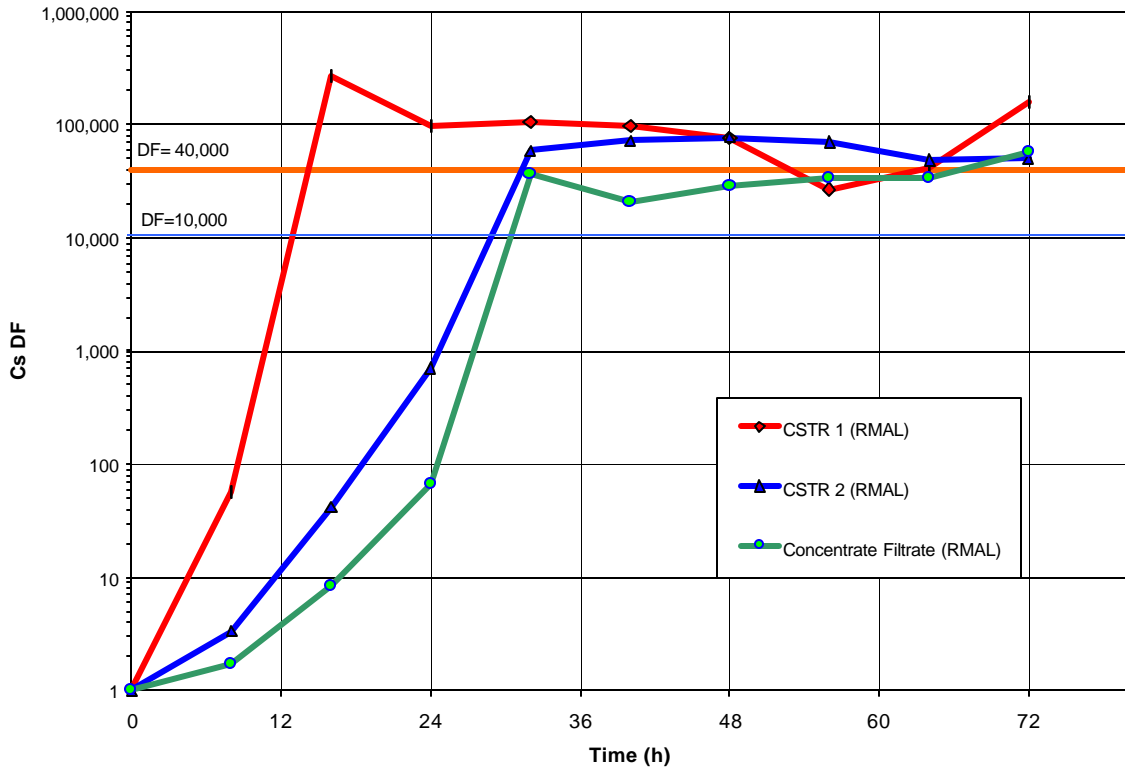


Fig. 9. Test 3 <sup>137</sup>Cs DF.

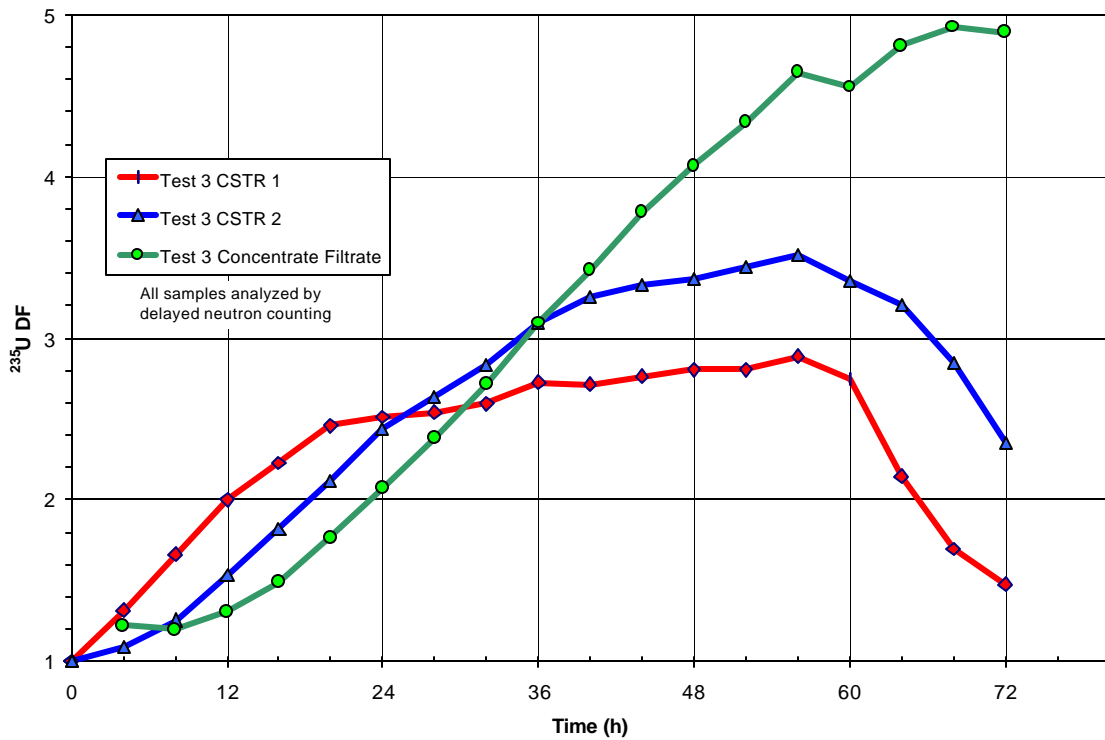


Fig. 10. Test 3 <sup>235</sup>U DF.

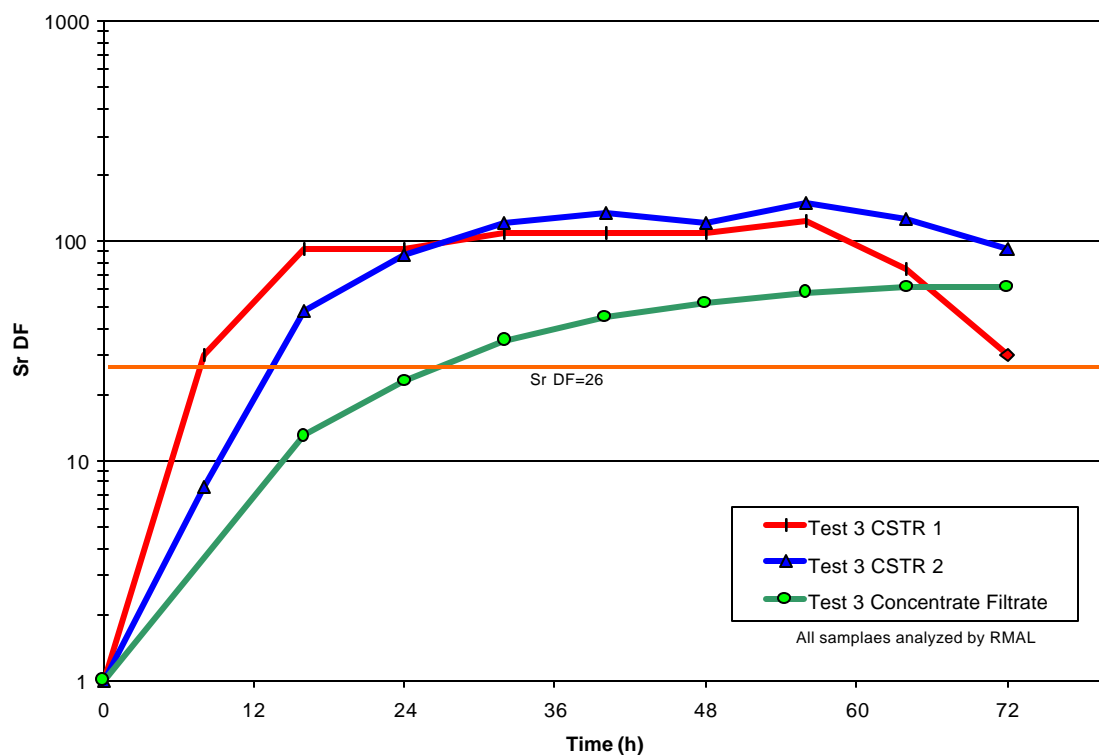


Fig. 11. Test 3  $^{85}\text{Sr}$  DF.

delivery system. The planned and actual flow summary is shown in Table 4. Afterward, both cross-flow filter systems were chemically cleaned with 2% oxalic acid solution, rinsed well with water, and then rinsed and stored in 1 M NaOH.

Table 4. Process performance for Test 3

	Planned volume (L)	Actual volume (L)	Average flow rates (mL/min)
Salt feed	92.88	93.02	21.53
Recycle wash water	29.64	29.78	6.89
NaTPB	2.938	2.980	0.69
Water + MST	8.94	8.84 <sup>a</sup>	2.05
Antifoam to CSTR 1	0.00673	0.00673	0.00156
Antifoam to CSTR 2	0.00673	0.00673	0.00156
Antifoam to SCT	0.01346	0.01346	0.00312
Filtrate from cross-flow filter	127.1	124.7 (plus ~2-L holdup in pipes and wash system vessels)	29.02
Slurry produced	6.5	6.5	

<sup>a</sup>Flow was interrupted during the last 6 to 8 h of MST/water feed due to an undetected breach in the feed tubing.

## 4.2 DETAILS OF SYSTEM CHEMISTRY

### 4.2.1 Cesium Decontamination of Test 3 Salt Feed

Figure 9 shows the DFs for cesium obtained during Test 3 in the CSTRs and the SCT filtrates based upon RMAL counting results. A  $DF_{Cs}$  of 40,000 was obtained in 14 h for CSTR 1 and thereafter ranged between 41,000 and 270,000. Only one sample at 56 h had a  $DF_{Cs}$  less than 40,000; it was 27,000. It took about 32 h to reach 40,000 in CSTR 2. Afterward, the  $DF_{Cs}$  ranged from 49,000 to 77,000. A  $DF_{Cs}$  of 37,000 was obtained in the SCT in about 32 h. It subsequently fluctuated between 21,000 and 59,000 and reached the highest value at the end of the test. Cesium concentrations obtained by both on-site and RMAL counting for all the system vessels agreed very well, as shown in Fig. 12. These data show a decrease in concentration of  $^{137}Cs$  as a function of test time in each of the vessel until the minimum cesium concentration is reached. In CSTRs 1 and 2, the concentration then slightly increases as steady state is approached. The RMAL data consistently showed slightly lower cesium concentrations for each filtrate sample. The long delay in obtaining  $DF_{Cs}$ , especially in the SCT filtrate, was the result of the method of starting the test (i.e., the vessels were filled with radioactively traced salt feed simulant before TPB was supplied). As TPB was added, the cesium and potassium reacted with TPB and precipitated. The soluble concentrations of these elements decreased until a state of equilibrium was established. To ensure a high DF, a mole ratio of TPB to potassium and cesium of 1.6 to 1 was used. Part of the delay in obtaining a high cesium DF in SCT filtrate was caused by the untreated salt feed retention in the 700-mL pressure vessel that was used for backpulsing the cross-flow filter. The cesium and potassium in the untreated salt feed in this vessel were slowly replaced by the treated filtrate.

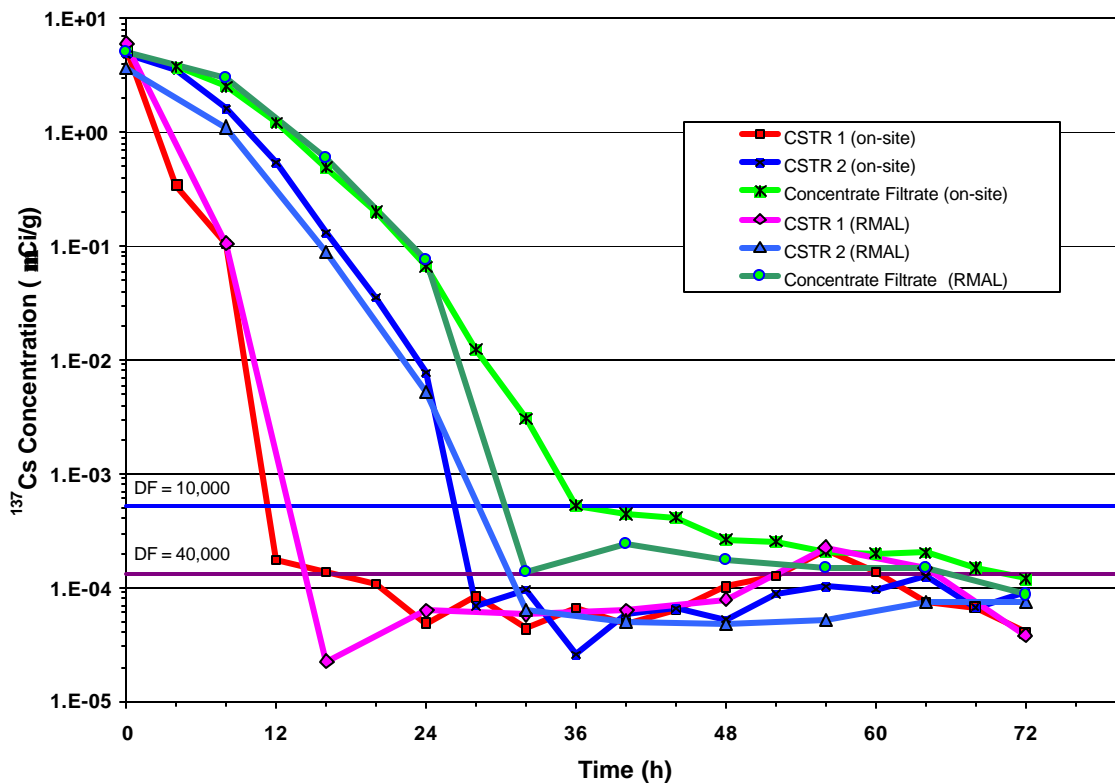


Fig. 12.  $^{137}Cs$  after CSTR 1, CSTR 2, and in the concentrate filtrate during Test 3.



Cesium removal data from Tests 1a and 2 was compared to the results of Test 3 to evaluate differences in cesium removal performance. Figures 13–15 show the cesium removal data, as determined by on-site gamma counting (only the first 72 hours of Test 2 are shown). All three tests were begun in the same way, with Tests 1a and 3 identical except for the higher agitator speeds and the antifoam addition in Test 3. Test 2 had the modified Enhanced Comprehensive Catalyst system added to the feed that contained a number of hydrous metal oxides that may have affected the rate of TPB precipitation of the potassium and cesium; otherwise, it was the same as Test 1a during startup. The results for the two CSTRs for Test 1a and Test 3 were similar in that a  $DF_{Cs}$  of  $>40,000$  was obtained at about the same operating time. Overall, judging by the data, the IITB52 antifoam used in Test 3 apparently had no impact on the cesium precipitation process.

#### 4.2.2 Strontium Decontamination of Test 3 Salt Feed

Figure 11, which shows RMAL (official) gamma counting data, indicates the DFs obtained for  $^{85}Sr$  removal by MST powder from the diluted salt feed in the CSTRs and SCT. Figure 16 shows the concentrations of both RMAL and on-site data for the CSTRs and SCT and the agreement is generally good. According to the RMAL counting data, it required about 24 h to obtain a DF of 30 for strontium in the SCT; the  $DF_{Sr}$  slowly increased to about 100 at the end of the test. A  $DF_{Sr}$  of 26 was needed to reach the WAC standard for saltstone. The reasons for the delay in obtaining the desired  $DF_{Sr}$  were similar to those that delayed reaching the desired  $DF_{Cs}$ . At the start of the test, the strontium-to-MST ratio was very high and the rate of sorption is somewhat slow. Moreover, the delay at the start is increased by the untreated salt feed in the 700 mL back-pulsing pressure vessel. The IITB52 did not appear to impede the sorption of strontium by the MST. These data show that a decrease was noted in the  $DF_{Sr}$  for CSTR 1 during the last +8 h of the test that occurred when the MST feed line to CSTR 1 came loose at a valve fitting. It was not detected until the filtrate samples taken during this time period were counted, alerting the operators to the problem. The balance data for the MST/water feed tank showed no loss of flow. The  $DF_{Sr}$  for CSTR 2 was also beginning to decrease at the end of the test.

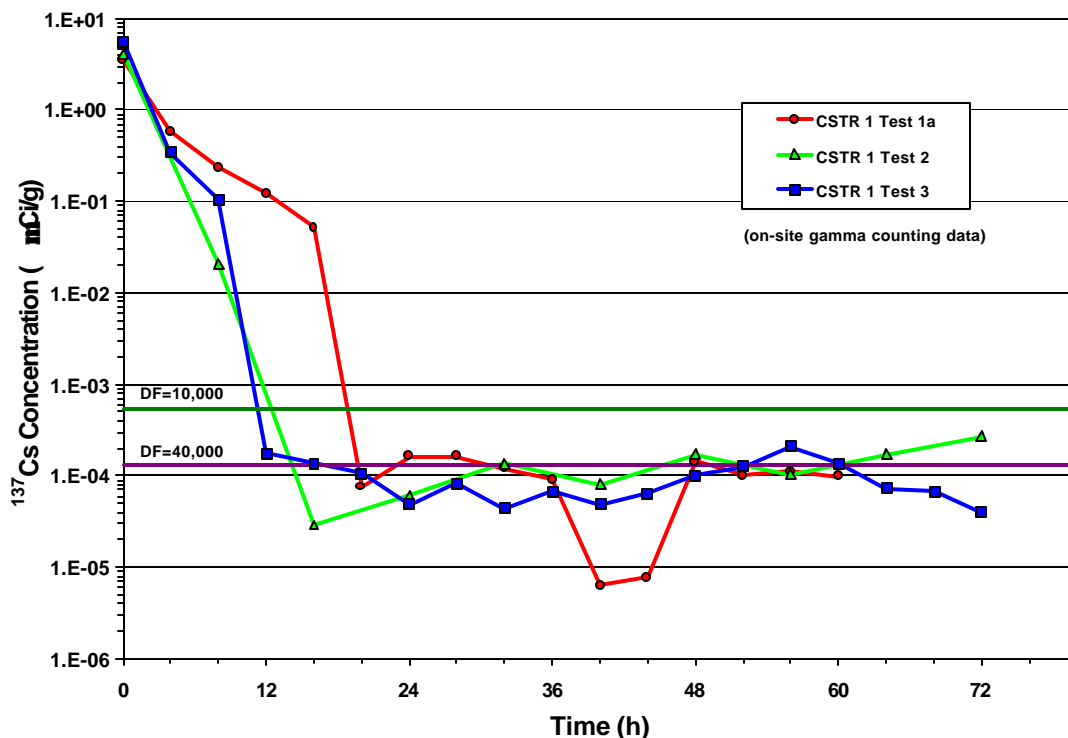


Fig. 13. Comparison of  $^{137}Cs$  removal for Tests 1, 2, and 3 for CSTR 1.

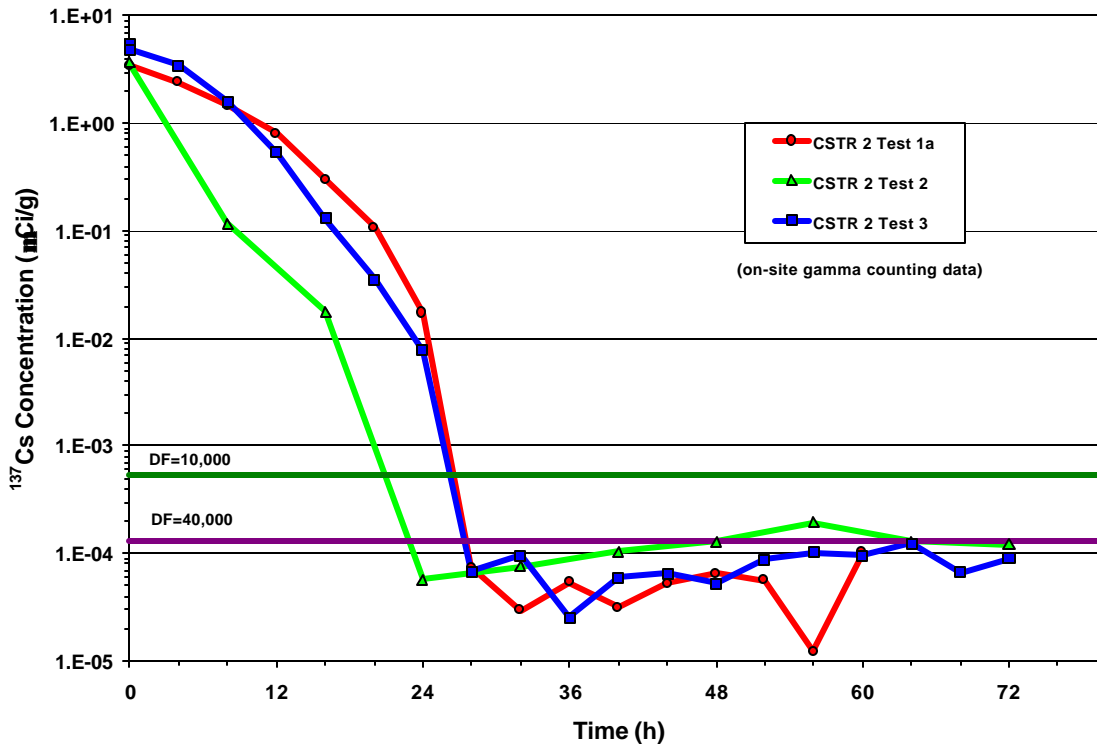


Fig. 14. Comparison of  $^{137}\text{Cs}$  removal for Tests 1, 2, and 3 for CSTR 2.

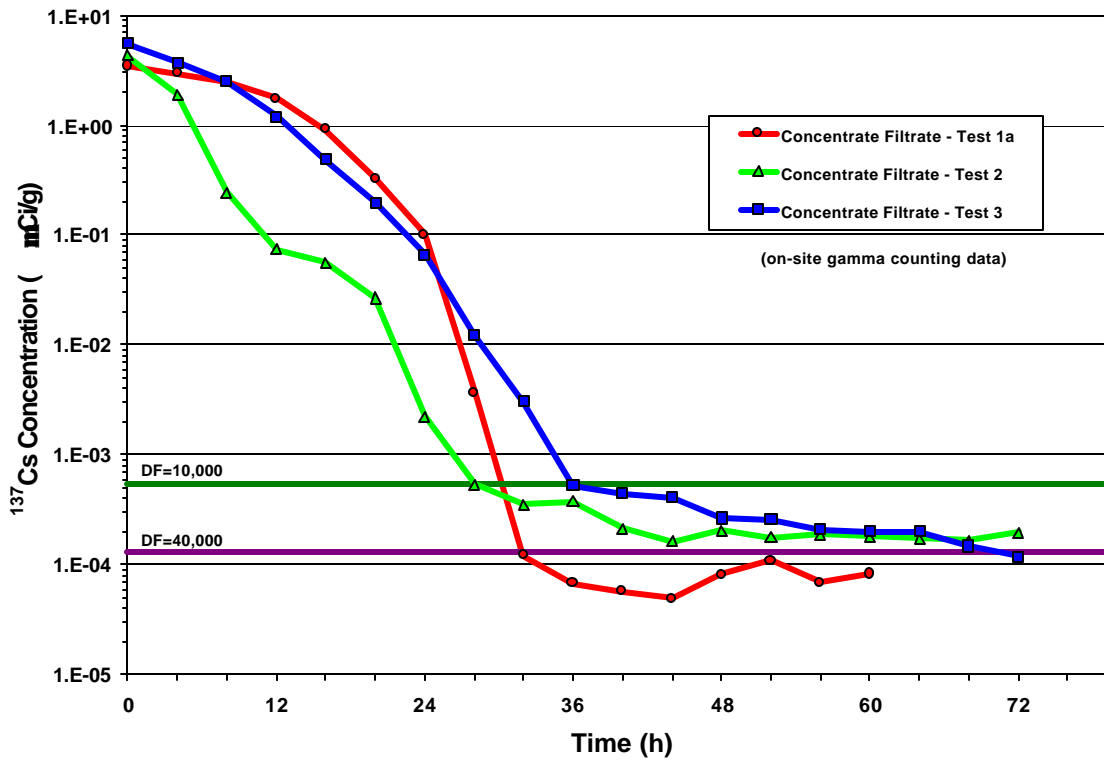


Fig. 15. Comparison of  $^{137}\text{Cs}$  removal for Tests 1, 2, and 3 for concentrate filtrate.

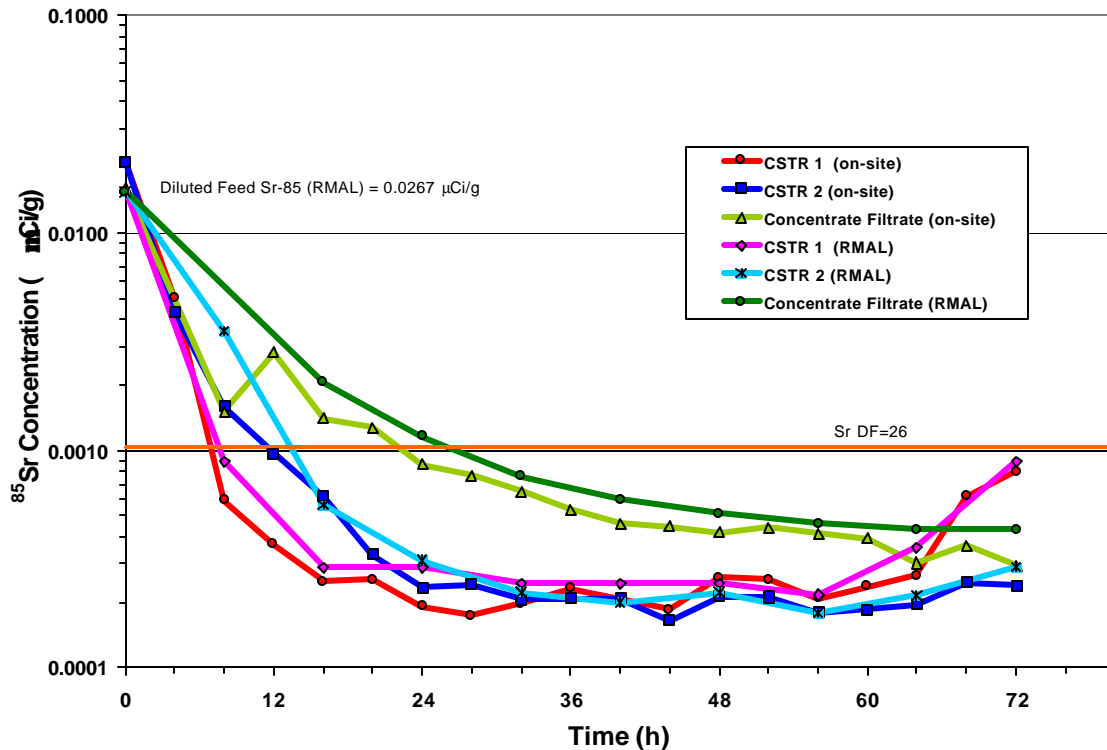


Fig. 16. Strontium concentrations during Test 3.

The RMAL strontium removal results for Test 1a and 3 for all the vessels are compared in Fig. 17, and the on-site results for individual vessels are provided in Figs. 18–20. There is fairly good agreement in the shape of the curves, but the percentage removal in Test 3 was greater. It is speculated that the increased mixer speed used in Test 3 contributed to the improved strontium removal. A decrease in  $DF_{Sr}$  for Test 3, as indicated by an increase in the strontium concentration, occurred during the last three or four sample periods in both CSTRs due to the inadvertent loss in MST feed. The stage effect of the three vessels is apparent for both strontium and uranium, as the concentration of both are progressively lower for CSTR 1, CSTR 2, and the SCT. The data indicates no interference of the IITB52 antifoam with strontium removal for this test.

#### 4.2.3 Uranium Decontamination of Test 3 Salt Feed

As measured by RMAL, a DF of 5 for uranium in the SCT was obtained at the end of the test, as shown in Fig. 10. This indicates that the other actinides (neptunium and plutonium), had they been present in the salt feed, probably would have been removed to meet WAC standards. A  $DF_U$  of 2 was reached after 12, 18, and 22 h in CSTR 1, CSTR 2, and the SCT, respectively. Later in the test, as steady-state conditions were approached, the concentrations of uranium in the three process vessels were aligned in accordance with the expected stagewise behavior of the process: the concentration in CSTR 1 was greater than CSTR 2, and the lowest uranium concentration was in the SCT. The data demonstrate the importance of residence time in the sorption of uranium by the MST. As with strontium, there was a reduction in uranium DF that is very noticeable for CSTR 1 in the 64- to 72-h time frame and for CSTR 2 in the 68- to 72-h time frame. For CSTR 2, the DF decreased from 3.5 to 2.5, and for CSTR 1, from 2.9 to 1.5. As mentioned above, the loss of MST flow to CSTR 1 caused the problem. Figure 21 compares the concentration of nonsorbed uranium in the vessels as a function of test time.

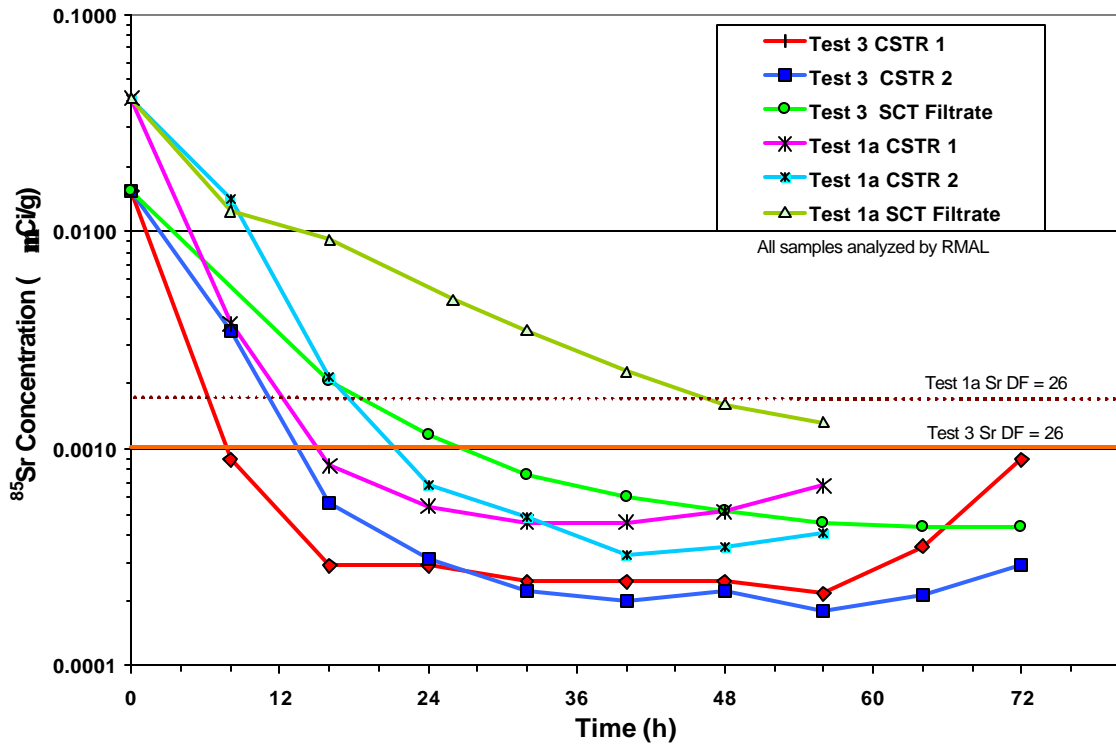


Fig. 17. Strontium concentrations during Tests 1a and 3.

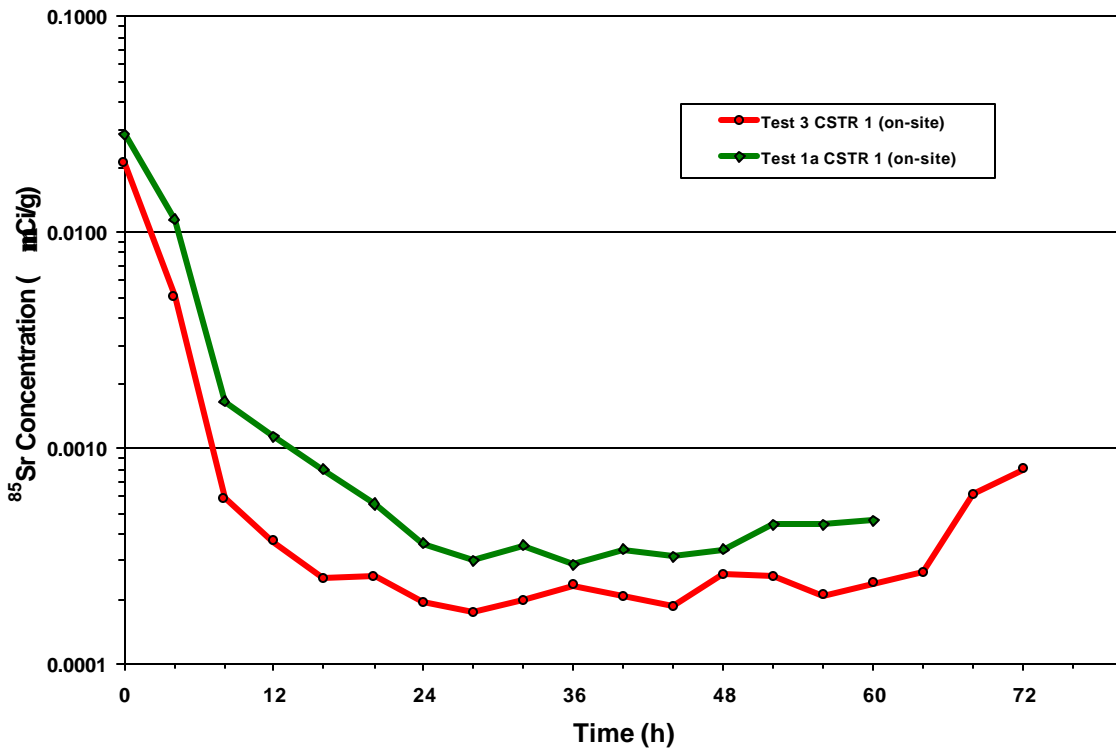


Fig. 18. Strontium concentrations after CSTR 1 in Tests 1a and 3.

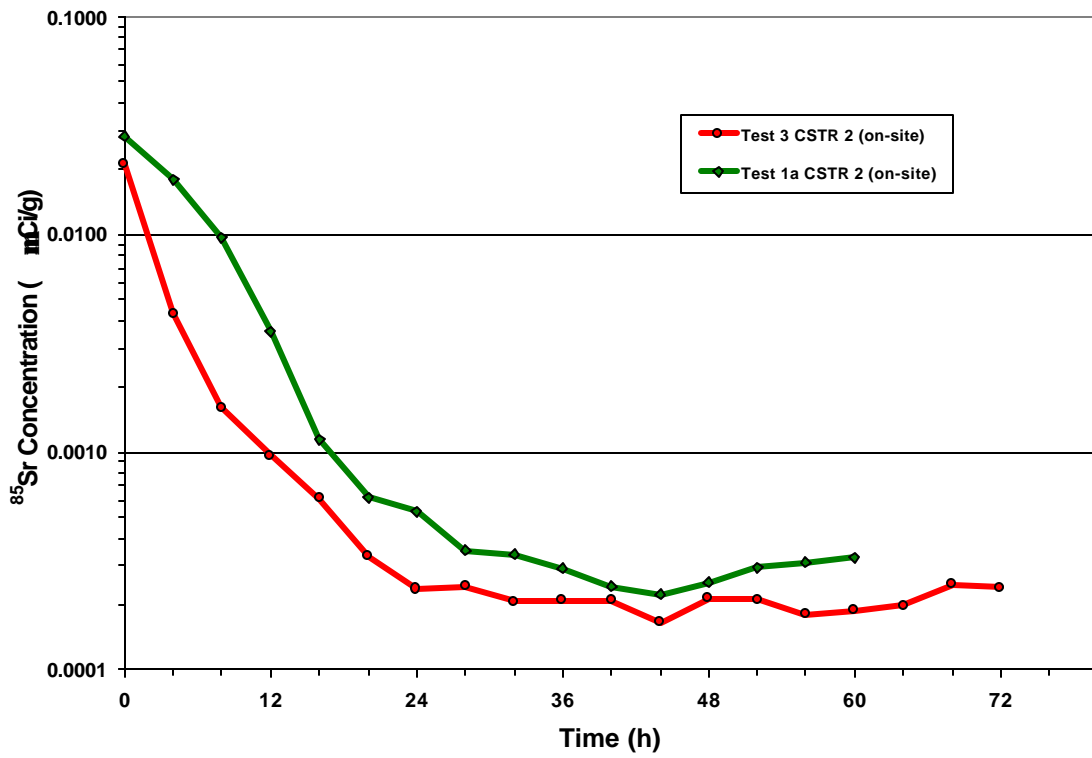


Fig. 19. Strontium concentrations after CSTR 2 in Tests 1 and 3.

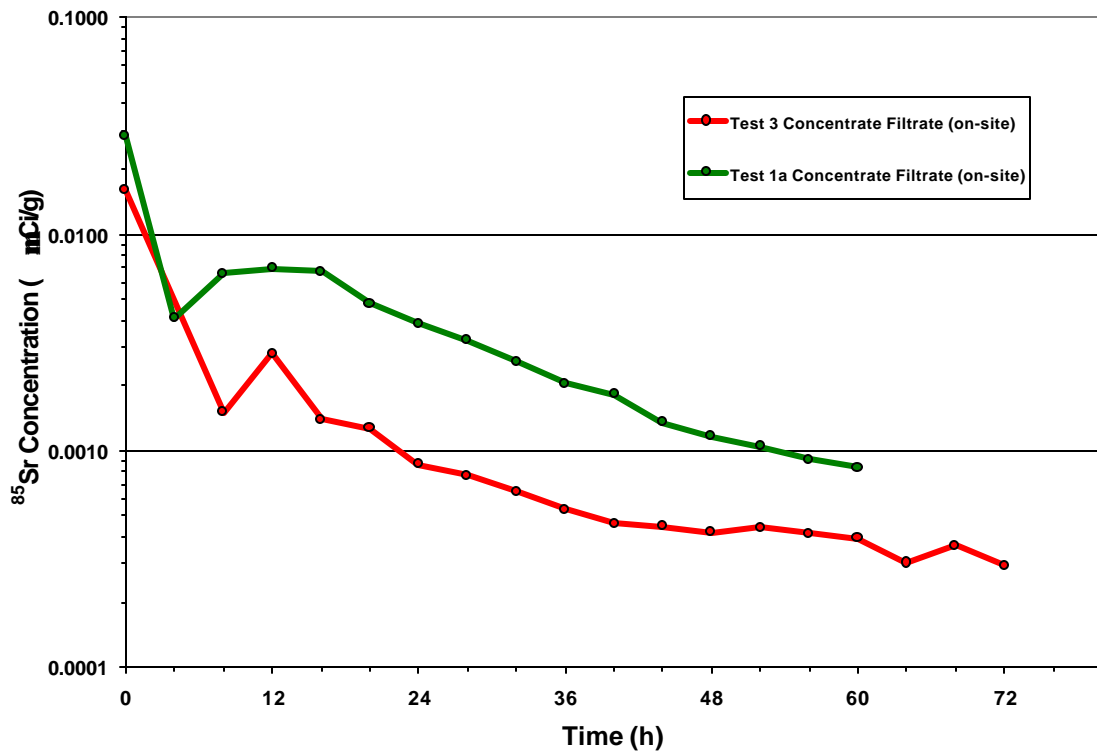


Fig. 20. Strontium concentrations in concentrate filtrate in Tests 1a and 3.

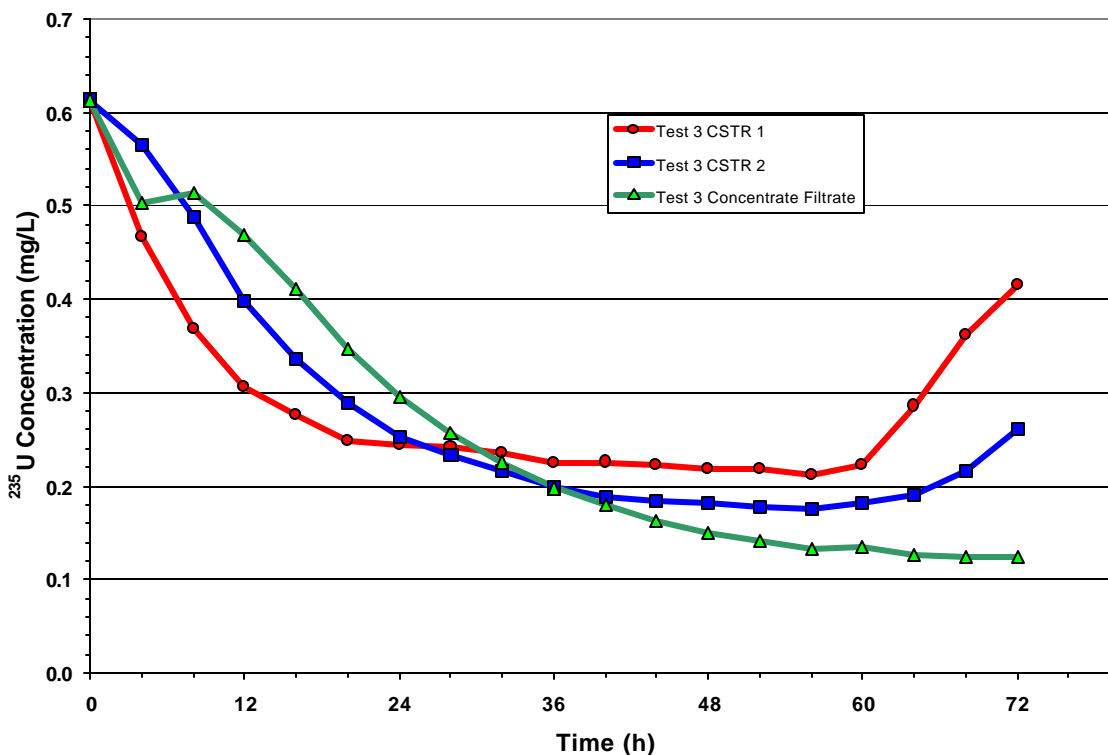


Fig. 21. Comparison of uranium concentrations during Test 3.

Figure 22 compares the uranium results obtained in Tests 3 and 1a. After 60-h testing, the Test 3 DFs were 5 in the SCT, 3.5 in CSTR 2, and 2.3 in CSTR 1—compared with DFs of 3, 2.5, and 2.3 for Test 1a. The removal was better in Test 3 even though the IITB52 was used. It is speculated that the increased mixer speed used in the process vessels in Test 3 enhanced the removal of uranium.

#### 4.2.4 HPLC Analyses for TPB and Decomposition Products

RMAL analysis of the samples from all three sample points showed less than the detectable amount of 1PB, 2PB, 3PB, and phenol throughout the test. The HPLC analyses showed that no measurable NaTPB decomposition occurred during the test (i.e., no decomposition products were detected). TPB analyses are given in Figs. 23–25. TPB follows the expected course, at first falling to undetectable levels as a soluble constituent in CSTR 1 as the bulk of the potassium and cesium was being precipitated from the diluted salt feed, and then increasing to levels of 100–150 ppm soluble NaTPB when the potassium concentration was reduced to about 200 ppm.

#### 4.2.5 ICP-AES Analyses for Potassium and Boron

Figures 23–25 show the RMAL results for the potassium, boron, and NaTPB for the three sample points. The potassium begins falling when TPB is added to CSTR 1. The boron analysis shows that the amount of boron is dependent on, and agrees with, the amount of soluble TPB present, as expected. If decomposition were taking place, the amount of boron would be greater than the amount associated with the soluble TPB.

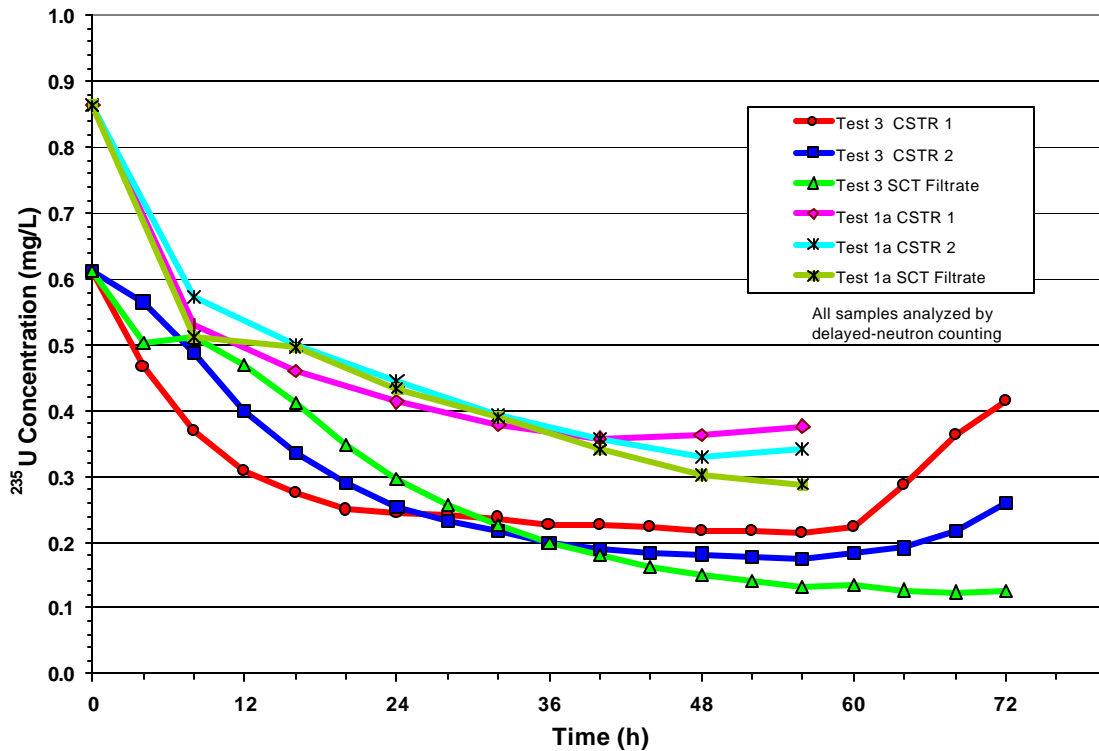


Fig. 22. Comparison of uranium concentrations during Tests 1a and 3.

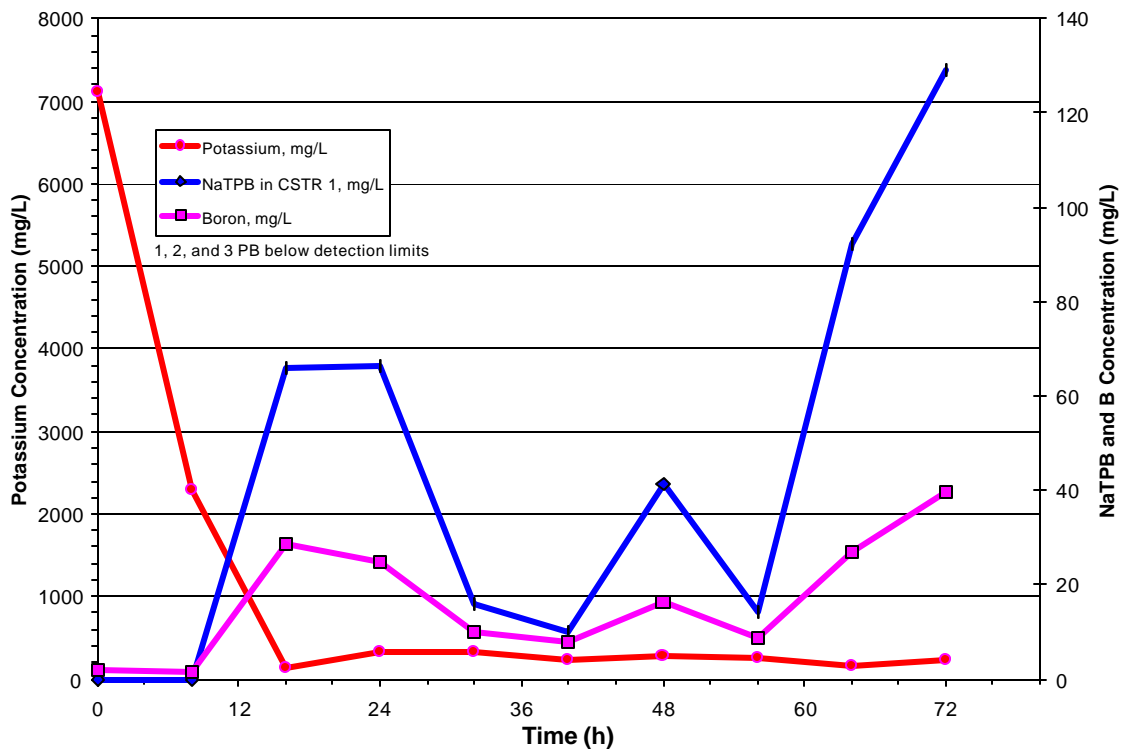
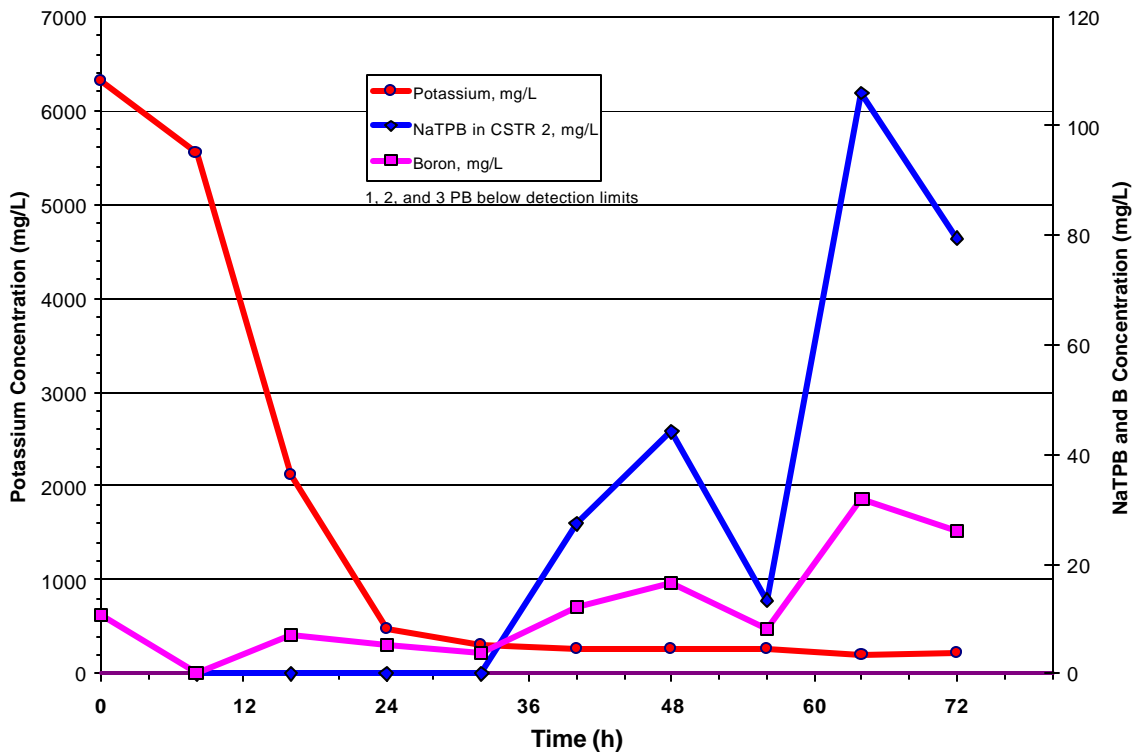
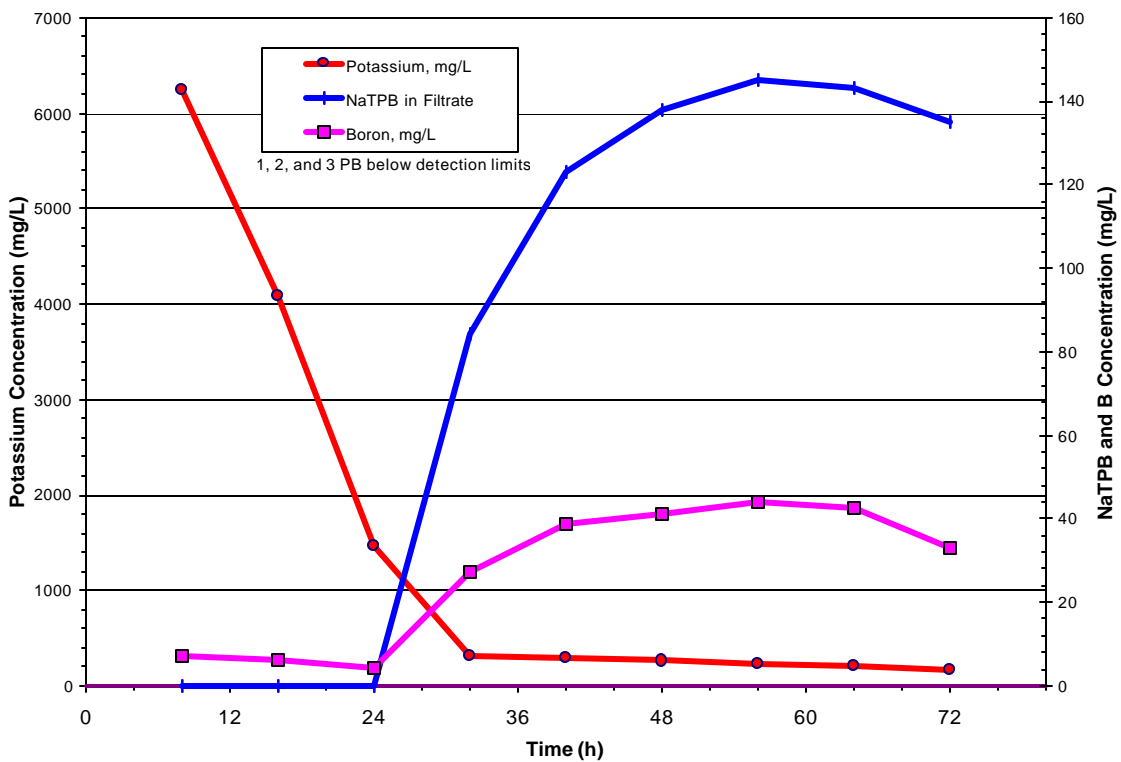


Fig. 23. Potassium, boron, and NaTPB after CSTR 1 during Test 3.



**Fig. 24. Potassium, boron, and NaTPB after CSTR 2 during Test 3.**



**Fig. 25. Potassium, boron, and NaTPB in concentrate filtrate during Test 3.**



## 4.2.6 Material Balances

The record of flow rate versus time for the process fluids used and produced in the CSTR system is shown in Figs. 26 and 27. Figure 26 shows the cumulative average flow rates for the several feeds. Additional information on the feed to the system is shown in Fig. 27. It shows the cumulative totals of each feed stream during the course of the test. The data indicates acceptable control of process feeds and smooth operation of the slurry concentrating system.

## 4.3 SUBSYSTEMS OPERATION

Process control for this test was sufficient to provide meaningful process performance data. Even with the high mixing speeds of 1200 rpm (shown to cause foam as noted in Sect. 2.1) in the process vessels (compared to the normal 600 rpm in previous tests in which no foam was seen), the process control in Test 3 was very good. Because the IITB52 antifoam performed well, foam formation was minimal. The solution density for the concentration system, as measured by the Coriolis mass flow measurement system, decreased in the normal manner during concentration from about 1.2 g/mL at the start to about 1.15 at the end of concentration. There was no significant reduction in density caused by the entrainment of foam in the slurry (foam would reduce the density by 0.2–0.5 g/mL and would be pressure sensitive). The radar probe in the salt feed tank functioned properly and provided good crosscheck measurements of the salt feed rate to CSTR 1 by comparing tank volume versus time to the instantaneous flow rate data. The differential pressure level probe (Bindicator) in the SCT operated with no problems. This level monitoring system was a significant improvement over the previous type of level sensor, a guided wave radar probe that plugged with slurry during operation and quit operating. The first four hours of slurry concentration were run using flow control, a less desirable method from a process control standpoint. When it was clear that the Bindicator monitor was operating reliably, the slurry concentration operation was switched to level control for the remainder of the test.

The larger trim in the filtrate-flow control valve in the cross-flow filtration loop (replacing the smaller trim used in the previous tests) provided stable and trouble-free operations at lower filtrate pressure (20 psi). This improvement also resulted in a lower heat load in the slurry concentrating system because the progressive-cavity filtration loop pump could be operated at a lower speed. The needed flow rates were obtained at 60 to 70% valve opening at 3- to 6-psi transmembrane pressure (pressure difference between the slurry concentrate and filtrate). Backpulsing the cross-flow filter to reduce transmembrane pressure was not necessary during the test. However, one back-pulse operation was performed after the run was completed to empty the backpulse pressure vessel.

Only a few difficulties were encountered at startup in achieving proper operation of all the feed systems. The first problem was caused by a loose fitting in the salt feed line that was revealed by air bubbles in the line while the salt feed was being pumped from the feed tanks in Hot Cell C to CSTR 1. This was corrected by replacing the plastic tubing of the salt feed pump (peristaltic type) and tightening the line fittings. When the NaTPB feed pump was started, no flow was seen (indicated by no change in the mass of the balance that the NaTPB tank sat on). It was assumed that the line was plugged at the cell wall or inside the cell. To correct the problem, the NaTPB line outside the hot cell was attached with a tee fitting to the simulated recycle wash water line at the wall of the hot cell. All the NaTPB was pumped into CSTR 1 via the recycle wash water line. It was later determined that the pumping difficulty was caused by an air bubble that became lodged in the pump head of the FMI positive displacement pump head used for the NaTPB.

A problem with the MST slurry feed occurred about 8–12 h before the test ended. Apparent settling of the slurry in the feed line caused overpressuring of the line. The excessive pressure caused the flexible feed line to breach at a location not clearly visible to the operators. The problem was not detected until the test was terminated and the valve at the hot cell wall was closed.

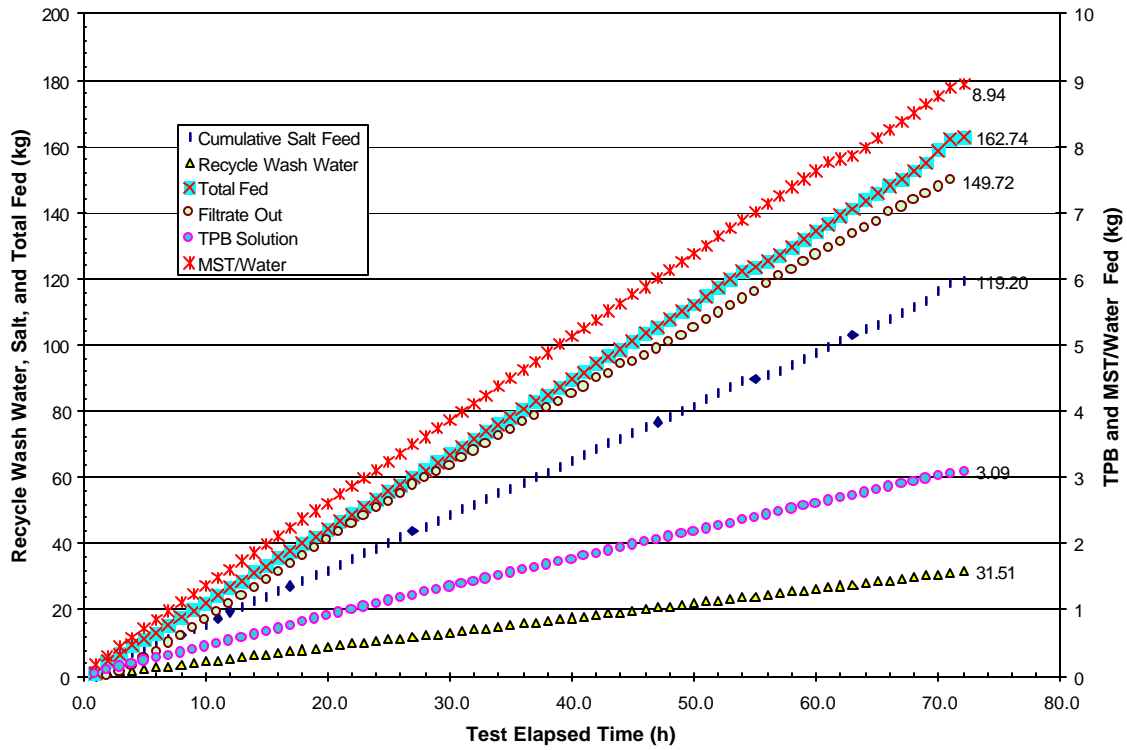


Fig. 26. Plot of the average feed rates of process fluids for Test 3.

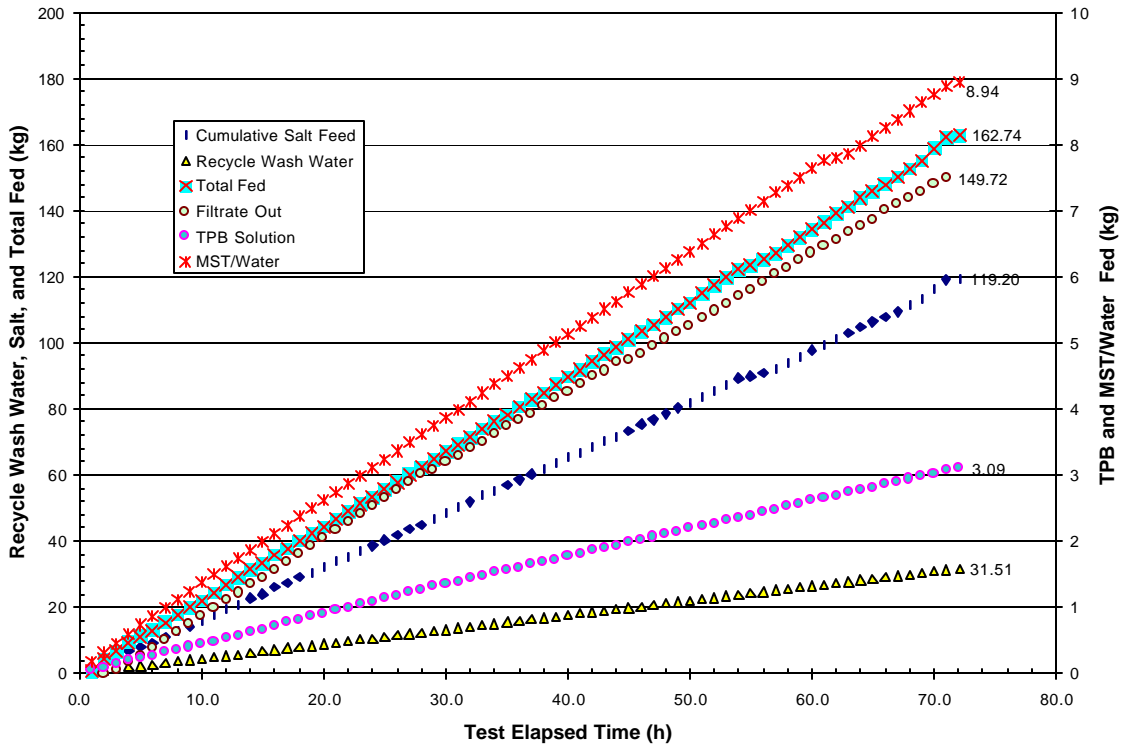


Fig. 27. Plot of the cumulative feed totals for materials fed to CSTR 1 during Test 3.

Analyses of  $^{85}\text{Sr}$  and  $^{235}\text{U}$  in the CSTR 1 filtrates showed decreases in DFs of those isotopes at the end of the test, which is consistent with loss of the needed MST concentration. The test was actually terminated 8 h earlier than planned due to the exhausted supply of simulant feed. The feed was exhausted somewhat prematurely because the dip tube for the feed tank did not reach the design depth in the tank. It was found in a later inspection of the tank that the dip tube was inadvertently bent upward during the movement of the feed system equipment from Hot Cell A to Hot Cell C. This premature shutdown did not impact the value of the test results in any way. Repairs were made prior to conducting additional tests. Aside from the feed problems previously noted, the overall equipment performance for Test 3 was very good and the feed rates and totals were all satisfactory.

#### 4.4 SLURRY WASHING

All the concentrated slurry produced during the test was transferred to the Slurry Washing Tank (SWT) and washed with deionized water to evaluate the ability to recover unreacted NaTPB. Deionized water was pumped into the SWT at a rate of 15 mL/min. The concentrated slurry was washed until the nitrite concentration of the cross-flow filter filtrate was about 0.0157 *M*, or 721 mg/L (see Table 5). The slurry solids were stirred at 1200 to 1250 rpm at 25°C while pumping IITB52 into the tank at a rate of 0.00312 mL/min. The vapor space of the SWT was purged with nitrogen gas during the washing process. Since the washing process was not conducted under closed-looped test conditions in which the slurry would be continuously washed, the solids were initially washed in three 8-h steps (~7.2 L of water during 8 h). This allowed for two overnight mixing periods, providing additional time for dissolution without increasing the volume of recycle water. For maximal dissolution of the NaTPB solids, the nitrite concentration of the slurry wash in the SWT should be about 0.065 *M* for the second overnight mixing period (the actual nitrite concentration was 0.078 *M* after 16 h and 0.031 *M* after 24 h of washing). A sample of the cross-flow filtrate was analyzed for NaTPB after the three 8-h steps of washing were complete. The nitrite had not reached 0.01 *M* after 24 h of washing; therefore, the washing was continued an additional 8 h when the nitrite concentration reached 0.0157 *M*. After further analysis was completed, the washwater was reserved for use in Test 4 initial startup operations, and additional NaTPB (9.9 g) was added to the 28.8 L of recycle wash water to obtain the needed concentration of 0.033 mol/L NaTPB for use in Test 4.

**Table 5. Results for Test 3 slurry washing at 15 mL/min**

Washing volume (L)	Washing time (h)	Recycle wash water sample ID	NO <sub>2</sub> (mg/L) <sup>a</sup>	NaTPB <sup>a</sup> (mg/L)
0	0	00T3W-1	20800	267
3.6	4	04T3W-2	14400	510
7.2	8	08T3W-3	9130	850
10.8	12	12T3W-4	6470	1300
14.4	16	16T3W-5	3630	1140
18.0	20	20T3W-6	1800	1120
21.6	24	24T3W-7	1440	948
25.2	28	28T3W-8	971	1440
<u>28.8</u>	<u>32</u>	<u>32T3W-9</u>	<u>721</u>	<u>1260</u>
Overall tank sample		36T3W-10	5740	1120

<sup>a</sup>Analyzed by RMAL.

The composite recycle wash water in the Recycle Wash Water Tank was thoroughly mixed and a sample taken for analysis. RMAL determined that the NaTPB was only 0.0033 mol/L. Because of the method of starting the test, in which the CSTRs were filled with the 4.7 M Na<sup>+</sup> water diluted salt feed (also 3 L in the SCT), a large fraction of the TPB was precipitated as potassium tetraphenylborate (KTPB), leaving only 0.33 mol of excess TPB in the slurry concentrating system available for water washing recovery. If all the TPB precipitated as NaTPB had dissolved in the 28.8 L of wash water, the concentration would have been 0.011 mol/L. This would have been only about 35% of the required 0.033 mol/L. About 0.095 mol of TPB was actually recovered, providing a concentration 0.0033 mol/L. This was 29% of the available TPB. The profiles of nitrate and NaTPB concentrations as a function of wash volume are shown in Fig. 28. The data indicates that NaTPB concentration reaches a consistent level in the range of 948–1440 mg/L (0.0028–0.0042 mol/L) in the washwater at a wash volume of 10 L and remains at this level during the entire wash cycle. This indicates that additional washing could recover additional quantities of NaTPB, but at the expense of generating washwater quantities greater than the amount that can be recycled efficiently. To obtain the desired recycle wash water TPB concentration of 0.033 mol/L, only 10 L of deionized water could have been used in the washing process.

Table 6 compared the washing results in Test 3 with those of the previous CSTR tests, which were also started up by first filling the CSTRs with the 4.7 M Na<sup>+</sup> water-diluted salt feed and adding about 3 L in the SCT. In Test 1a, the concentrated slurry was washed at a rate of 10 mL/min, and 12.3 L of wash water was collected. The concentrated slurry in Test 1a was mixed at 600 to 700 rpm at 25°C and the antifoam IITB52 was not used. Twelve per cent of the available TPB (0.048 mol) was recovered. However, less wash water was used and the TPB concentration was 0.0047 mol/L, which was higher than the TPB concentration of 0.0033 mol/L for the Test 3 wash water. None of the water washes in Test 2 (also stirred at 600–700 rpm at 25°C with no antifoam IITB52) recovered more than 20% of the available TPB. The Test 2, Cycle #4 batch may have recovered more TPB if the wash had continued for the same volume as the Test 3 wash, possibly twice as much as shown. If Test 3 wash had been stopped at about 14 L, the amount of TPB recovered would have been about half, as seen in Fig. 28. Although the recoveries were generally low in every case, the trends of these data in Table 5 show the importance of the washing parameters. These low concentrations of NaTPB in the wash water may be indicative of inefficient mixing that caused poor mass transfer and NaTPB dissolution rates at the lower stirrer speeds. Based on the results in Table 6, it is probable that the higher mixer speed used in test 3 could have enhanced TPB recovery.

The solubilities of the TPB precipitates of Na, K, and cesium are the basis for the utility of the STTP.<sup>6,7</sup> TPB dissolution behavior was studied in laboratory-scale tests with IITB52 antifoam at Savannah River.<sup>8</sup> Several tests of TPB dissolution were conducted in which slurries that had been produced by precipitation of a standard 4.7 M sodium simulant with 0.0125 M potassium were concentrated to 10 wt % using a cross-flow filter system. Laboratory batch mixing tests were performed for one slurry without antifoam and three others with three different antifoams, one of which was IITB52. The results showed that about 60% of the excess TPB was recovered, and a maximum concentration of NaTPB in the wash water was 0.041 M. When IITB52 was used, only 12.7% of the excess NaTPB was removed and the maximum NaTPB concentration was 0.0058 M. These large differences between the tests with and without antifoam were not observed in the 20-L system in Tests 1a and 3 in which the slurry produced for washing had a slightly higher concentration of KTPB (less NaTPB to dissolve) due to the method of startup.

**Table 6. NaTPB recovery in the slurry washing process in CSTR tests**

Test No.	Wash cycle	Time (h)	Wash rate (mL/min)	Available TPB <sup>a</sup> (mol)	RWW <sup>b</sup> (L)	TPB recovered (mol)	TPB <sup>c</sup> (mol/L)	Percent recovered <sup>d</sup>
1a	1 <sup>e</sup>	16	10	0.48	12.3	0.058	0.0047	12
2	1	31	10	0.86	18.6	0.072	0.0039	9
2	2	28	15	0.63	25	0.115	0.0046	18
2	3	43	8	0.76	20.6	0.148	0.0072	19
2	4	44	5	0.69	13.2	0.138	0.0105	20
3	1 <sup>f</sup>	32	15	0.33	28.8	0.095	0.0033	29

<sup>a</sup>Moles of excess TPB as NaTPB in batch available to be dissolved by water washing.

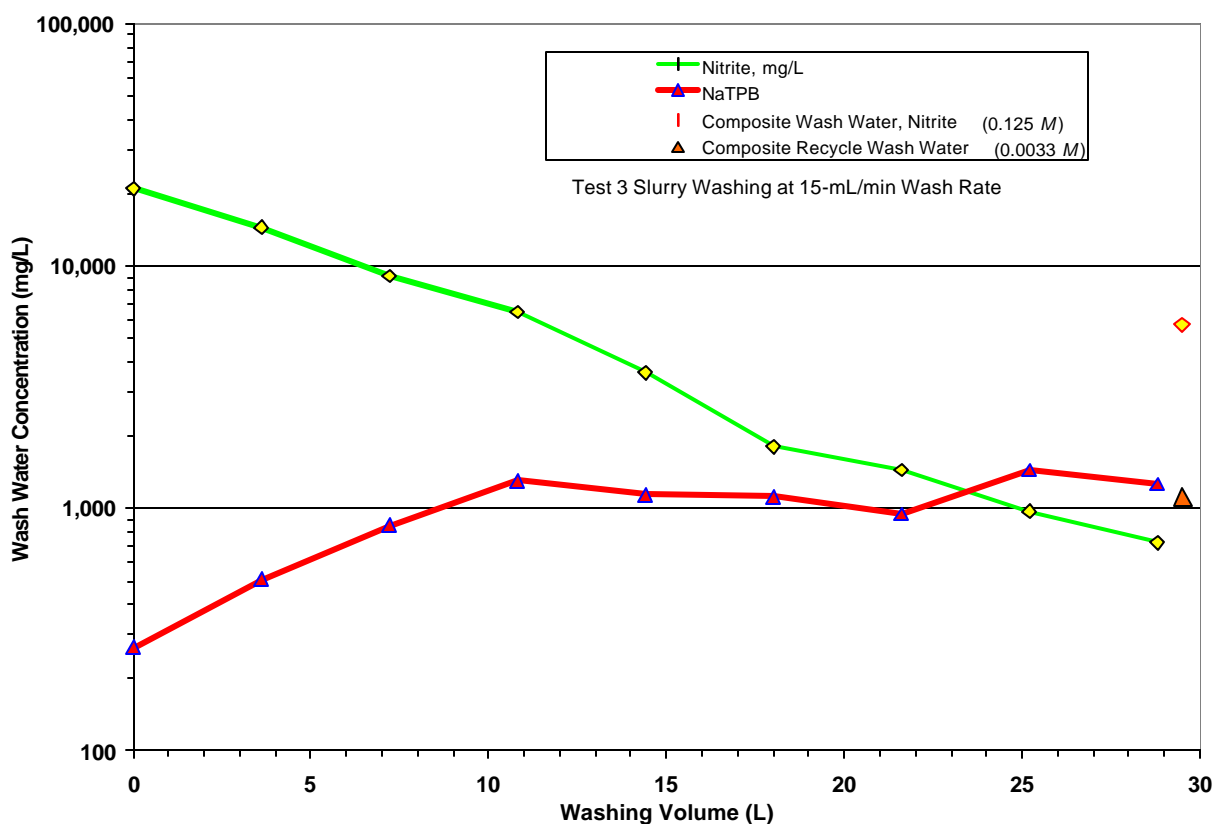
<sup>b</sup>RWW = recycle wash water.

<sup>c</sup>Testing goal was to obtain enough TBP in recycle wash water to provide a 60% excess (~0.033 mol/L).

<sup>d</sup>(mol TPB recovered)/(mol available TPB in batch).

<sup>e</sup>Recycle wash water used in Test 2.

<sup>f</sup>Recycle wash water to be used in Test 4.



**Fig. 28. Nitrite and TPB concentrations during Test 3 slurry washing.**

## 5. CONCLUSIONS

Test 3 was a 72-h test with approximately nine volume turnovers that was conducted with the objectives of determining the effectiveness of an improved antifoam (IITB52) in minimizing foam formation and to determine if its presence had a deleterious effect on the removal of cesium, strontium, and uranium from simulated waste. To more aggressively test the antifoam, the slurry in each vessel was mixed at 1200 to 1250 rpm while maintaining the temperature at 25°C. This rate of mixing was found to cause heavy foam formation in a control test that was conducted prior to Test 3. There was no evidence of excessive foam formation in Test 3, as indicated by the lack of foam in the analytical samples of slurries taken, by the absence of foam in the clear plastic slurry transfer lines between the vessels, and by lack of any significant decrease the slurry density measurements in the slurry concentration loop. More than 92 L of radioactive, simulated waste was treated, ~135 L of decontaminated filtrate was collected, and ~6.5 L >9 wt % slurry concentrate was collected and washed.

The decontamination goal of obtaining a DF of 40,000 or 99.998% removal was exceeded in all process vessels, indicating that the IITB52 did not impact cesium removal performance. A  $DF_{Cs}$  of 37,000 was obtained in the SCT in about 32 h. It subsequently fluctuated between 21,000 and 59,000 and was the highest at the end of the test. A  $DF_{Cs}$  of 40,000 was obtained in 14 h CSTR 1 and afterward ranged between 41,000 and 270,000. It took about 32 h for the DF to reach 40,000 in CSTR 2, and for the remainder of the test, the DF ranged from 49,000 to 77,000. In comparing the cesium concentration versus time profiles for previous tests where the antifoam was not used, there was no clear evidence that the rate of cesium removal was adversely impacted by the antifoam used in this test.

The IITB52 had no impact on the sorption of strontium and uranium by the MST. A DF of 30 was obtained for strontium in the SCT in about 24 h, and near the end of the test, it was 100. A  $DF_{Sr}$  of 26 was needed to reach the WAC standard for saltstone. A DF of 5 for uranium in the SCT was obtained at the end of the test. This is indicative that the other actinides (neptunium and plutonium) would achieve DFs that met the WAC standards for saltstone.

At the end of Test 3, the concentrated slurry in the concentrating system was transferred to the SWT and washed with deionized water to evaluate the ability to recover unreacted NaTPB. The slurry solids were stirred at 1200 to 1250 rpm at 25°C while adding wash water at 15 mL/min and pumping IITB52 into the tank at a rate of 0.00312 mL/min. A total of 28.8 L of wash water that had a concentration of 0.0033 mol/L (29% of the available TPB) was collected. This recovery was actually better than the recovery in Test 1a and Test 2 in which only 12% and 18–20% of the available TPB was recovered, respectively, and no antifoam was used. It does not appear that the antifoam interfered with the dissolution of the precipitated NaTPB. The higher mixing rate in Test 3 may have contributed to the larger fraction of TPB that was recovered. There is an obvious problem in recovering available TPB by water washing that needs to be experimentally examined more closely.

Problems with the operation of the equipment during the test were minimal, especially in the slurry concentration cross-flow filter system. Process modifications and improved control and data acquisition systems improved the overall operating performance of the equipment. The primary problems involved the feed systems. However, these process upsets were dealt with quickly and efficiently by the operating personnel without significant impact on the overall process or the value of the operating data.

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